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Abstract

Soil fertility depletion (mainly N, P and carbon) has been described as the single most important constraint to food security in West Africa. Over half of the African population is rural and directly dependent on locally grown crops. Further, 28% of the population is chronically hungry and over half of people are living on less than US\$ 1 per day as a result of soil fertility depletion.

Soil organic carbon (SOC) is simultaneously a source and sink for nutrients and plays a vital role in soil fertility maintenance. In most parts of West Africa agro-ecosystems (except the forest zone), the soils are inherently low in SOC. The low SOC content is due to the low shoot and root growth of crops and natural vegetation, the rapid turnover rates of organic material as a result of high soil temperatures and fauna activity particularly termites and the low soil clay content. With kaolinite as the main clay type, the cation exchange capacity of the soils in this region, often less than 1 cmol kg⁻¹, depends heavily on the SOC. There is a rapid decline of SOC levels with continuous cultivation. For the

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23 sandy soils, average annual losses may be as high as 4.7% whereas with sandy loam soils,
24 losses are lower, with an average of 2%. To maintain food production for a rapidly
25 growing population application of mineral fertilizers and the effective recycling of
26 organic amendments such as crop residues and manures are essential.

27 Crop residue application as surface mulch can play an important role in the maintenance
28 of SOC levels and productivity through increasing recycling of mineral nutrients,
29 increasing fertilizer use efficiency, and improving soil physical and chemical properties
30 and decreasing soil erosion. However, organic materials available for mulching are scarce
31 due to a low overall production levels of biomass in the region as well as their
32 competitive use as fodder, construction material and cooking fuel. Animal manure has
33 similar role as residue mulching for the maintenance of soil productivity but it will
34 require between 10 and 40 ha of dry season grazing and between 3 and 10 ha of
35 rangeland of wet season grazing to maintain yields on one hectare of cropland. The
36 potential of manure to maintain SOC levels and maintain crop production is thus limited
37 by the number of animals and the size and quality of the rangeland. The potential
38 livestock transfer of nutrients in West Africa is 2.5 kg N and 0.6 kg P per hectare of
39 cropland.

40 Scarcity of organic matter calls for alternative options to increase its availability for
41 improvement of SOC stock. Firstly, the application of mineral fertilizer is a prerequisite
42 for more crop residues at the farm level and the maintenance of soil organic carbon in
43 West African agro-ecosystems and therefore most research should focus on the
44 improvement of nutrient use efficiency in order to offer to the smallholder farmers cost-
45 effective mineral fertilizer recommendations. Secondly, recent success story on

46 increasing crop production and SOC at the farm level is the use of the dual purpose grain
47 legumes having ability to derive a large proportion of their N from biological N fixation,
48 a low N harvest and substantial production of both grain and biomass. Legume residues
49 can be used for improvement of soil organic carbon through litter fall, or for feeding
50 livestock with the resultant manure being returned to the crop fields.

51 In the decision support system for organic matter management, recommendations for
52 appropriate use of organic material was made based on their resource quality, expressed
53 as a function of N, polyphenol and lignin content. High quality organic materials release
54 a high proportion of their N quickly. The impact of organic resource quality on SOC is
55 less clear. Low quality organic resources contain substantial amounts of soluble
56 polyphenols and lignins that may affect the longer-term decomposition dynamics and
57 contribute to the build up of SOC. Future research needs to focus more on whether the
58 organic resource quality concept is also useful for predicting different degrees of
59 stabilization of applied organic C in one or more of the organic matter pools.

60

60 1) *Introduction*

61 Over half of the African population is rural, and directly dependent on locally grown
62 crops or foods harvested from the immediate environment. The growth rate for cereals
63 grain yield is about 1% while population growth is about 3%. During the last 35 years,
64 per capita cereals production has decreased from 150 to 130 kg/person, whereas in Asia
65 and Latin America an increase from about 200 to 250 kg/person have been observed.
66 Both labor and land productivity in Africa are among the lowest in the world. Per capita
67 food production in Africa has been declining over the past two decades, contrary to the
68 global trend. Annual cereal deficit in sub-Saharan Africa amounts to 100 million tons and
69 the food gap (requirements minus production) is widening. Food imports increased by
70 about 185% between 1974 and 1990 while food aid increased by 295%. The average
71 African consumes only about 87% of the calories needed for a healthy and productive
72 life. Sixteen percent (16%) of Africa's current arable land base is so eroded that
73 agriculturally it cannot be useful any longer. In addition to this, 70% of deforestation is
74 caused by farmers who in their quest for food have no incentive to ponder about long-term
75 environmental consequences.

76 Soil fertility depletion (mainly N, P and carbon) has been described as the single most
77 important constraint to food security in West Africa. The Sudano-Sahelian zone of West
78 Africa is the home of the world poorest people; 90% of who live in villages and gain their
79 livelihood from subsistence agriculture (Bationo and Buerkert, 2001). Per capita food
80 production has declined significantly over the past three decades. According to FAO,
81 total food production in Sahelian countries grew by an impressive 70% from 1961 to

82 1996, but it lagged behind as the population doubled causing per capita food production
83 to decline by approximately 30% over the same period.

84 Increasing population pressure has decreased the availability of arable land and it is no
85 longer feasible to use extended fallow periods to restore soil fertility. The fallow period
86 which would have restored soil fertility and organic carbon is reduced to lengths that
87 cannot regenerate soil productivity leading to systems non-sustainability (Nandwa, 2001).
88 High population densities have necessitated the cultivation of marginal lands that are
89 prone to erosion hence further environmental degradation through soil erosion and
90 nutrient mining. As a result, the increase in yield has been more due to land expansion
91 than to crop improvement potential. The 7.6% yield increase of yam in West Africa was
92 mainly due to an area increase of 7.2% and only 0.4% due to improvement in crop
93 productivity itself (Table 1).

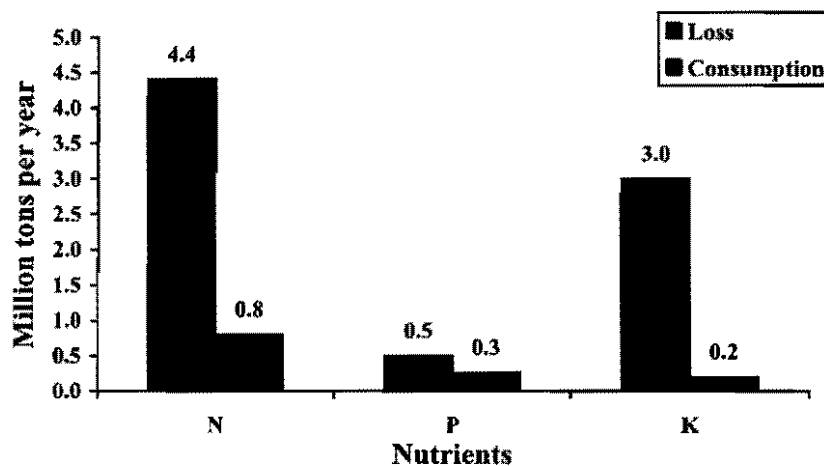
94 **Table 1:** Percentage annual increase in crop yield due to land expansion and crop
95 improvement potential in West Africa

Crops	Area (%)/year	Productivity (%)/year	Production (%)/Year
Cassava	2.6	0.7	3.3
Maize	0.8	0.2	1.0
Yam	7.2	0.4	7.6
Cowpea	7.6	-1.1	6.5
Soybean	-0.1	4.8	4.7
Plantain	1.9	0.0	2.0

96 **Based on three-year average for 1988-1990 and 1998-2000. Source: www.fao.org.**

97

98 Removal of crop residues from the fields, coupled with lower rate of macronutrient
99 application compared to losses, has contributed to negative nutrient balances (Stoorvogel
100 and Smaling, 1990). For nitrogen as an example, whereas 4.4 million tons is lost per year,
101 only 0.8 million tons is applied (Bationo et al., 2004) (**Figure 1**). Shepherd and Soule
102 (1998) reported a negative carbon balance of 400 kg ha⁻¹ yr⁻¹ for farmers with low
103 resource endowment whereas positive balance of 190 kg ha⁻¹ yr⁻¹ was reported on fields
104 of farmers with high resource endowment (**Table 2**). Additionally, low and erratic
105 rainfall, high ambient soil and air temperatures, inherent poor soil fertility, low water
106 holding capacities and degraded soil structure lead to low crop productivity in this
107 environment. Consequently, the present farming systems are not sustainable (Bationo and
108 Buerkert, 2001).



109

110 **Figure 1:** Nutrient losses versus application rate in Africa

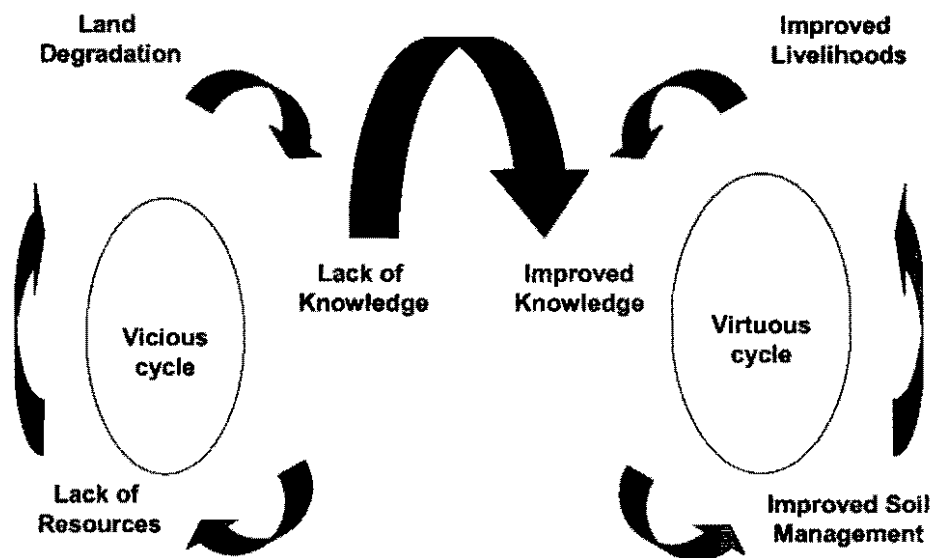
111 **Table 2:** Soil organic carbon balance, soil erosion, farm return and household income at
 112 different farm resource endowment levels

Variable	Units	Farm resource endowment		
		Low	Medium	High
Soil C balance	Kg ha ⁻¹ yr ⁻¹	-400	-318	190
Soil erosion	T ha ⁻¹ yr ⁻¹	5.6	5.5	2.1
Farm returns	\$ yr ⁻¹	3	70	545
Household income	\$ yr ⁻¹	454	1036	3127

113 Source: Shepherd and Soule (1998)

114 Transforming agriculture in West Africa agro-ecosystems and expanding its production
 115 capacity are prerequisites for alleviating rural poverty, household food deficits and
 116 environmental exploitation (Bationo et al., 2004). Reverting the declining trend in
 117 agricultural productivity and preserving the environment for present and future
 118 generations in West Africa must begin with soil fertility restoration and maintenance
 119 (Bationo et al., 1996). Soil fertility is closely linked to soil organic matter, whose status
 120 depends on biomass input and management, mineralization, leaching and erosion (Roose
 121 and Barthes, 2001; Nandwa, 2001). It is well recognized that soil organic matter increases
 122 structure stability, resistance to rainfall impact, rate of infiltration and faunal activities
 123 (Roose and Barthes, 2001). Optimum management of the soil resource for provision of
 124 goods and services requires the optimum management of organic resources, mineral
 125 inputs and the SOC pool (Vanlauwe, 2004). The importance of SOC has increased
 126 interest and research on its build up in the soil-plant system with current emphasis on
 127 conservation tillage. SOC can play an important role and its maintenance is an effective
 128 mechanism to combat land degradation and increase future food production.

129 The challenge in increasing SOC content is to embrace the holistic strategy of Integrated
130 Soil Fertility Management (ISFM) that puts into consideration the biophysical and socio-
131 economic constraints faced by the farmer community. The implementation of the ISFM
132 strategy will break the vicious cycle responsible for land degradation, food insecurity and
133 poverty in West Africa agro-ecosystems through improved knowledge of soil
134 management and the capacity of farmers to invest in improved soil management
135 technologies (Figure 2).



136

137 **Figure 2.** The vicious and virtuous cycles of land degradation and soil fertility
138 management

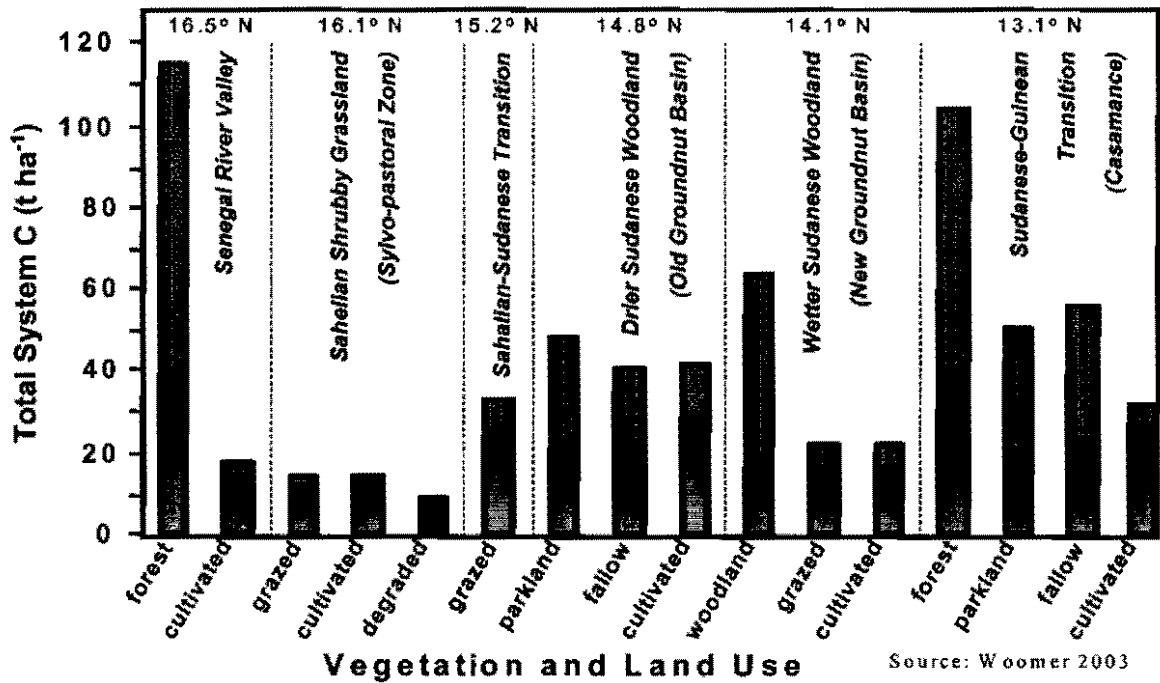
139 This paper will discuss first the status of soil organic carbon at agro-ecosystem and farm
140 level followed by the factors affecting SOC and functions of SOC before discussing the

141 effects of soil and crop management on SOC and concluding on the future research
142 challenges with emphasis on SOC quantity and quality.

143 *2) Soil organic carbon status at agro-ecosystem and farm level*

144 Soil organic carbon is an index of sustainable land management (Woomer et al., 1994,
145 Nandwa, 2001) and is critical in determining response to N and P fertilization. There is
146 however no clear agreement on the level of SOC below which response to N and P
147 fertilization does not occur. For example, while Berger et al. (1987) reported such level to
148 be 3.5 mg kg⁻¹ in the northern Guinean zone, Bationo et al. (1998) found very strong
149 response to mineral fertilizer at SOC levels as low as 1.7 mg kg⁻¹.

150 Total system carbon in different vegetation and land use types indicates that forests,
151 woodland and parkland have the highest total and aboveground carbon contents (**Figure**
152 **3**) demonstrating potential for carbon sequestration. For example, total system carbon in
153 the Senegal River valley was 115 t ha⁻¹ in the forest zone and only 18 t ha⁻¹ when the land
154 was under cultivation. Cultivated systems have reduced carbon contents due to reduced
155 tree cover and increased mineralization due to surface disturbance. Windmeijer and
156 Andriess (1993) found levels of SOC for equatorial forest, Guinea savanna and Sudan
157 savanna to be 24.5, 11.7 and 3.3 g kg⁻¹ respectively and showed positive correlation with
158 both N and P (**Table 3**).



159

160 **Figure 3.** Total system carbon in different agro-ecosystems and land use in West Africa

161 **Table 3:** Carbon stocks and other fertility indicators of granitic soils in different agro-
 162 ecological zones in West Africa

AEZ	pH (H ₂ O)	OC (g kg ⁻¹)	Total N (g kg ⁻¹)	Total P (mg kg ⁻¹)
Equatorial Forest	5.3	24.5	1.6	628
Guinea Savanna	5.7	11.7	1.39	392
Sudan Savanna	6.8	3.3	0.49	287

163 Source: Windmeijer and Andriess (1993)

164 SOC levels across fields on-farm show steep gradients resulting from long-term site-
 165 specific soil management by the farmer. According to Prudencio (1993), SOC status of
 166 various fields within a farm in Burkina Faso showed great variations with home gardens
 167 (located near the homestead) having 11-22 g kg⁻¹ (**Table 4**), village field (at intermediate
 168 distance) 5-10 g kg⁻¹ and bush field (furthest) having only 2-5 g kg⁻¹. Usually, closer fields

169 are supplied with more organic inputs as compared to distant fields due to the labour
 170 factor. Manu et al. (1991) found that SOC contents were highly correlated with total N
 171 ($r=0.97$) indicating that in the predominant agro-pastoral systems without application of
 172 mineral N, N nutrition of crops largely depend on the maintenance of SOC levels.

173 **Table 4:** Carbon stocks of different subsystems in a typical upland farm in the Sudan-
 174 savanna zone

AEZ	pH (H ₂ O)	OC (g kg ⁻¹)	Total N (g kg ⁻¹)	Available P (mg kg ⁻¹)	Exchangeable K (mmol kg ⁻¹)
Home garden	6.7-8.3	11-22	0.9-1.8	20-220	4.0-24
Village field	5.7-7.0	5-10	0.5-0.9	13-16	4.0-11
Bush Field	5.7-6.2	2-5	0.2-0.5	5-16	0.6-1

175 Source: Prudencio, (1993)

176 3) Factors affecting SOC

177 Clay and silt play an important role in the stabilization of organic compounds and small
 178 variations in topsoil texture could have large effects on SOC (Bationo and Buerkert,
 179 2001). In this context, a survey of West African soils (Manu et al., 1991) indicated that
 180 for the soils investigated cation exchange capacity (CEC) depended directly more on
 181 SOC ($r = 0.86$) than to soil clay content ($r = 0.46$) (Table 5). De Ridder and Van Keulen
 182 (1990) found a difference of 1 g kg⁻¹ in SOC to result in a difference of 0.25 cmol kg⁻¹ for
 183 soil CEC.

184 **Table 5:** Correlation (r) between selected soil (0-20 cm) fertility parameters and average
 185 annual rainfall

	Ca	CEC	SOC	Total N	Clay	Rainfall
pH KCl	0.62***	0.64***	0.65***	0.62***	-0.02	0.25**
Ca		0.98***	0.88***	0.92***	0.36***	0.31***
CEC			0.86***	0.91***	0.40***	0.36***
SOC				0.97***	0.46***	0.42***
Total N					0.44***	0.34***
Clay						0.40***

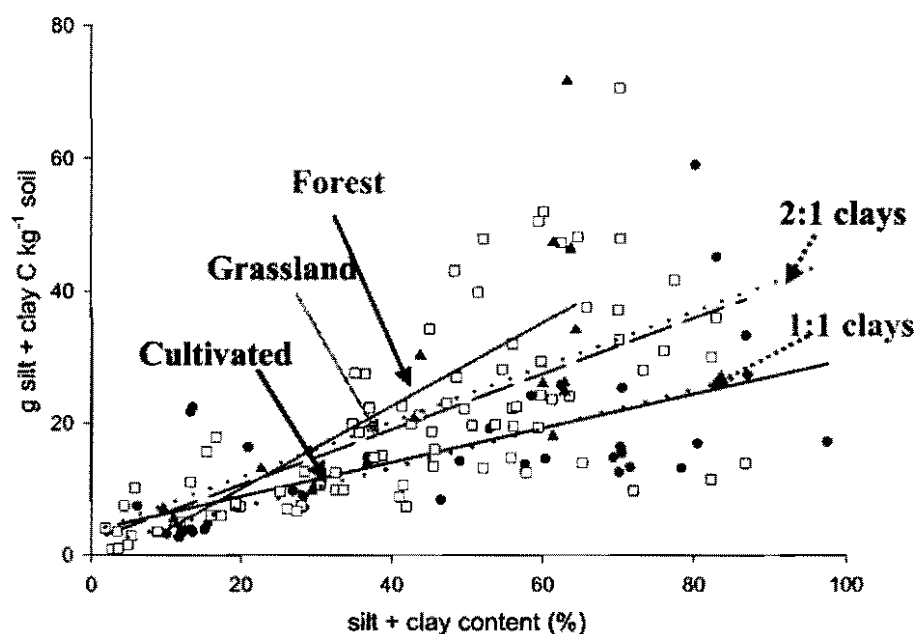
186 ** and *** indicate statistical significance at the 0.05 and 0.001 level, respectively.

187 Source: Manu et al. (1991)

188

189 **Figure 4** shows the relationship between silt and clay associated carbon and soil texture
 190 in different ecosystems and reflects the capacity of soil to preserve C based on its silt and
 191 clay particles. Carbon content and status in the soil is closely associated with clay and silt
 192 contents and clay type, which influences the stabilization of organic carbon. Aggregates
 193 physically protect SOC through formation of barriers between microbes and enzymes and
 194 their substrates thereby controlling microbial turnover (Six et al., 2002b).

195



Source: Six et al., 2002b

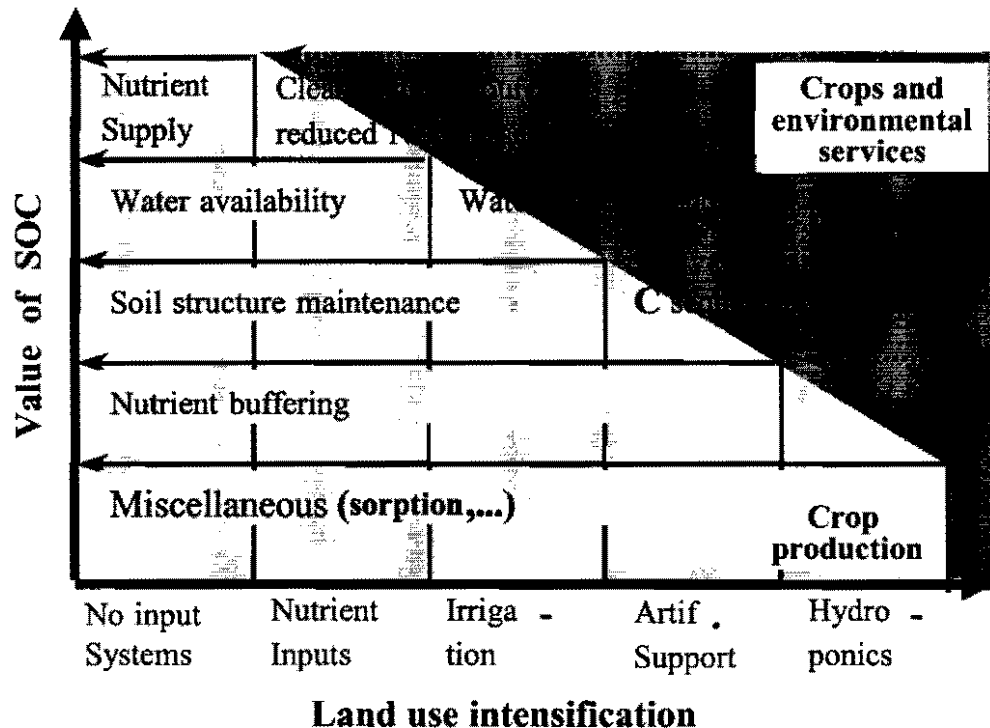
196 **Figure 4.** Relationship between silt+clay content (0-20 μm) and silt+clay associated
 197 carbon for different systems

198 **4) Functions of SOC**

199 SOC plays an important role in supplying plant nutrients, enhancing cation exchange
 200 capacity, improving soil aggregation and water retention and supporting soil biological
 201 activity (Dudal and Deckers, 1993). Although it has been difficult to quantify the effects
 202 of SOC on crop and ecosystem productivity (*Ibid*) results from experiments in some
 203 African countries already indicate favourable responses due to SOC.

204 Soil organic matter is not only a major regulator of various processes underlying the
 205 supply of nutrients and the creation of a favourable environment for plant growth but also
 206 regulates various processes governing the creation of soil-based environmental services

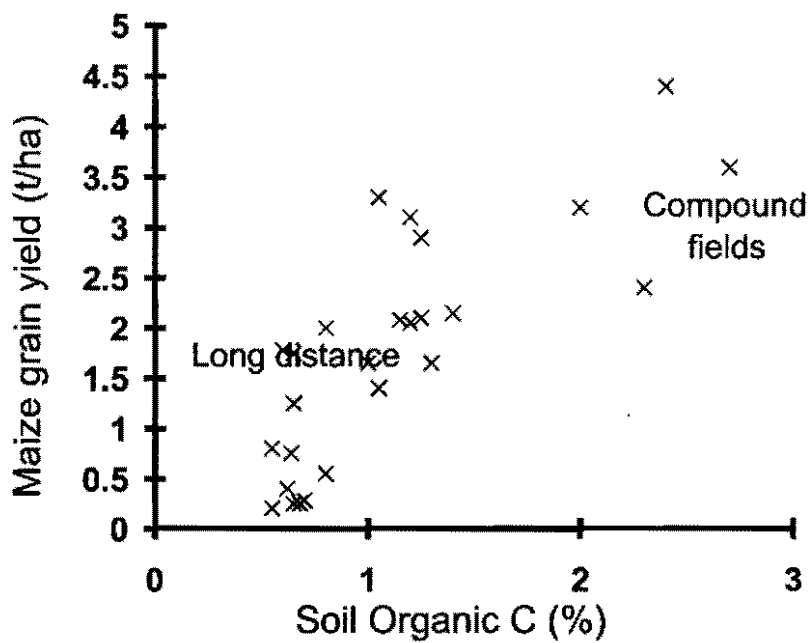
207 (Figure 5) (Vanlauwe, 2004). Therefore, SOC plays a vital role in crop production and
 208 environmental services.



209 **Figure 5.** Different functions of SOC and their regulation at different land intensification
 210 systems (Adapted from Vanlauwe, 2004)

211 **a) Crop production**

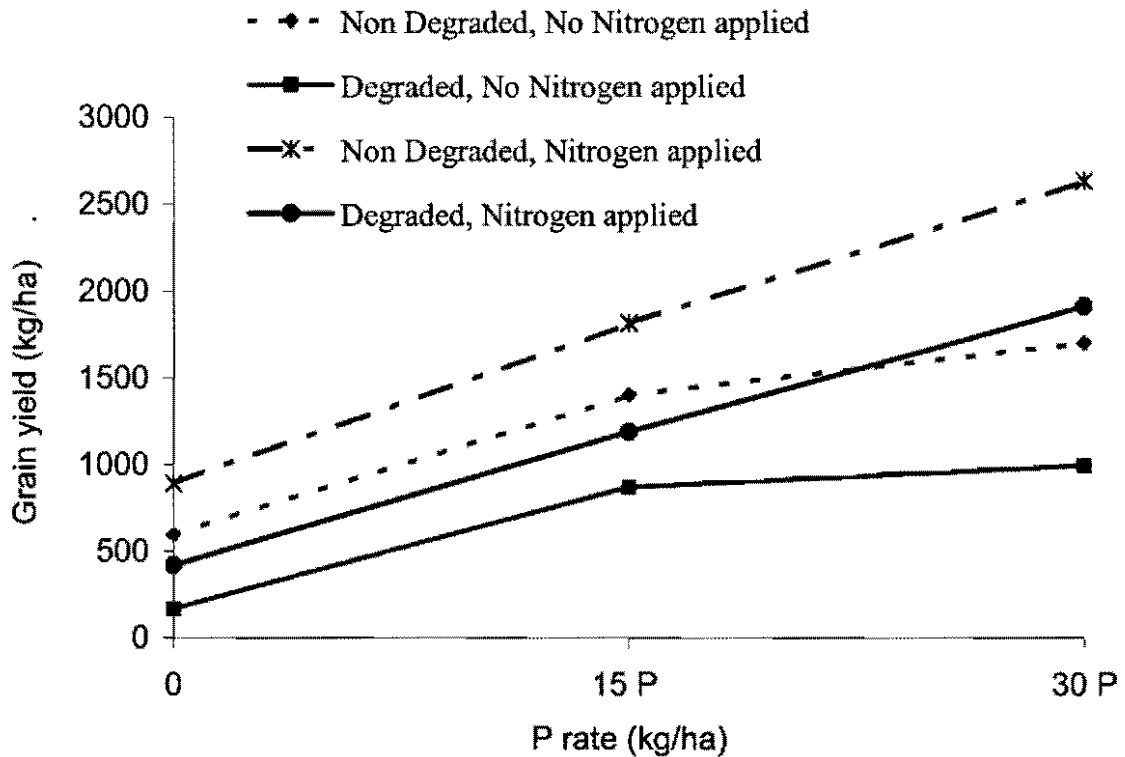
212 As already indicated, there is a steep gradient in SOC status between a field at the farm
 213 level scale caused by long-term site-specific soil management by farmers (Table 4). As
 214 shown in Figure 6, high SOC status in the homestead fields is observed to relate
 215 positively with crop yields (Figure 6). This is due to multiple factors of production
 216 affected by SOC content (Swift and Woomer, 1993).



218

219 **Figure 6.** Relationship between SOC content and maize grain yield for distant and
 220 compound fields in Northern Nigeria. **Source: Carsky et al. (1998)**

221 Over a period of four years in the Sahel, pearl millet yields on homestead fields with
 222 higher organic carbon were always significantly higher than yields in the bush fields
 223 lower in SOC content (**Figure 7**).



224

225 Figure 7. Effect of SOC content on millet production in Karabedji, Niger in 2002

226 Several scientists have reported the effect of organic amendments on crop yield increases
 227 partly due to effects of SOC (Bationo and Mokwunye, 1991; Bationo et al., 1995;
 228 Mokwunye, 1980; Pichot et al., 1981; Pieri, 1986; de Ridder and van Keulen, 1990;
 229 Abdulahi and Lombin, 1978; Powell, 1986 and Bationo et al., 1998). Research results
 230 from long-term field experiments in the West African agro-ecosystems showed that the
 231 use of mineral fertilizers without recycling of organic materials resulted in higher yields,
 232 but this increase was not sustainable (Jones, 1976; Bationo and Mokwunye, 1991).

233 As a result of the higher organic carbon content in mulched plots, Bationo et al. (1993)
 234 reported a large positive and additive effect of crop residue and mineral fertilizer
 235 application on pearl millet yield (Table 6). Over the duration of the study, grain yield in
 236 control plots (no fertilizer, no crop residue) were low and steadily declined. This
 237 indicated that the potential for continuous millet production on these soils is very limited
 238 in the absence of soil amendments. Except for the drought year in 1984, fertilizer
 239 application resulted in an approximately tenfold increase compared to the control. Since
 240 the P fixation of the sandy soils of the Sahel is low (Mokwunye et al., 1986) and residual
 241 effects of P-fertilizer application are evident even after three years, the use of P-fertilizer
 242 has important implications for sustainable soil management. The availability of cheap P
 243 fertilizers to small farmers may induce them to cultivate less land more intensively
 244 thereby leaving more area under fallow or pasture. This, in turn, would decrease the
 245 negative effects of wind and water erosion on the soil productivity.

246 **Table 6:** Effect of crop residue and fertilizer on pearl millet grain and stover yields at
 247 Sadore, Niger.

Treatment	Grain yields (kg ha ⁻¹)				Stover yield (kg ha ⁻¹)			
	1983	1984	1985	1986	1983	1984	1985	1986
Control	280	215	160	75	NA	900	1100	1030
Crop residue (no fertilizer)	400	370	770	745	NA	1175	2950	2880
Fertilizer (no crop residue)	1040	460	1030	815	NA	1175	3540	3420
Crop residue plus fertilizer	1210	390	1940	1530	NA	1300	6650	5690
LSD _{0.05}	260	210	180	200		530	650	870

248

249 *b) Ecosystem services*

250 The relevance of SOC in regulating soil fertility decreases as natural capital is being
251 replaced by manufactured or financial capital with increasing land use intensification
252 (**Figure 5**) (Vanlauwe, 2004).

253 Carbon sequestration has gained momentum in the recent decade and the amount of
254 carbon in a system is a good measure of sustainability. The current importance on this
255 subject is because carbon lost from these systems contributes significantly to atmospheric
256 change, particularly CO₂ concentration (Woomer and Palm, 1998). Estimates of carbon
257 stocks within different land management and cropping systems are an important element
258 in the design of land use systems that protect or sequester carbon (*Ibid*). Tropical
259 countries offer a large potential of carbon sequestration through reforestation and
260 improvement of degraded agroecosystems (Dixon et al., 1993). The limited studies in
261 small hold agricultural farms in Africa have already illustrated significant increases in
262 system carbon and productivity through organic-inorganic resources management
263 (Woomer et al. 1997 and Roose and Barthes, 2001). The data in **Table 6** indicates that
264 cereal biomass production can be increased by over 5 times from 1,030 to 5,690 kg ha⁻¹
265 when both crop residue and fertilizer are used in production. It is obvious that the
266 application of crop residue and fertilizer will increase both below and above ground
267 carbon sequestration.

268 Soil organic carbon plays an important role in ensuring good health of the soil
269 environment and is critical in providing needed ecosystem services (**Figure 5**). A higher
270 content of SOC will result in a higher Fertilizer Use Efficiency (FUE) (**Figure 7**). For

271 example, as a consequence of higher SOC content in the homestead fields, fertilizer use
 272 efficiency was higher as compared to the bush field. With application of 26 kg P ha⁻¹ in
 273 Karabedji Niger in 2000, P use efficiency was 42% in degraded site as compared to 79%
 274 in the non-degraded site (**Figure 7**). Comparative data of P FUE with and without crop
 275 residues mulch application in the Sahel clearly indicate a better fertilizer use efficiency
 276 with organic amendments which improve SOC (**Table 7**).

277 **Table 7.** Increase in incremental millet grain and stover yield due to fertiliser application
 278 in Sadore, Niger

Year	Treatment	Fertilizer effect (kg per kg P applied)	
		Grain	Stover
1983	Crop Residues	-	-
	Fertilizer	59 ¹	NA
	Crop Residues + Fertilizer	72 ²	NA
1984	Crop Residues	-	-
	Fertilizer	34	21
	Crop Residues + Fertilizer	14	31
1985	Crop Residues	-	-
	Fertilizer	67	188
	Crop Residues + Fertilizer	137	427
1986	Crop Residues	-	-
	Fertilizer	57	184
	Crop Residues + Fertilizer	112	359

279 1. Calculated as (Yield Fertilizer - Yield Control) / P applied

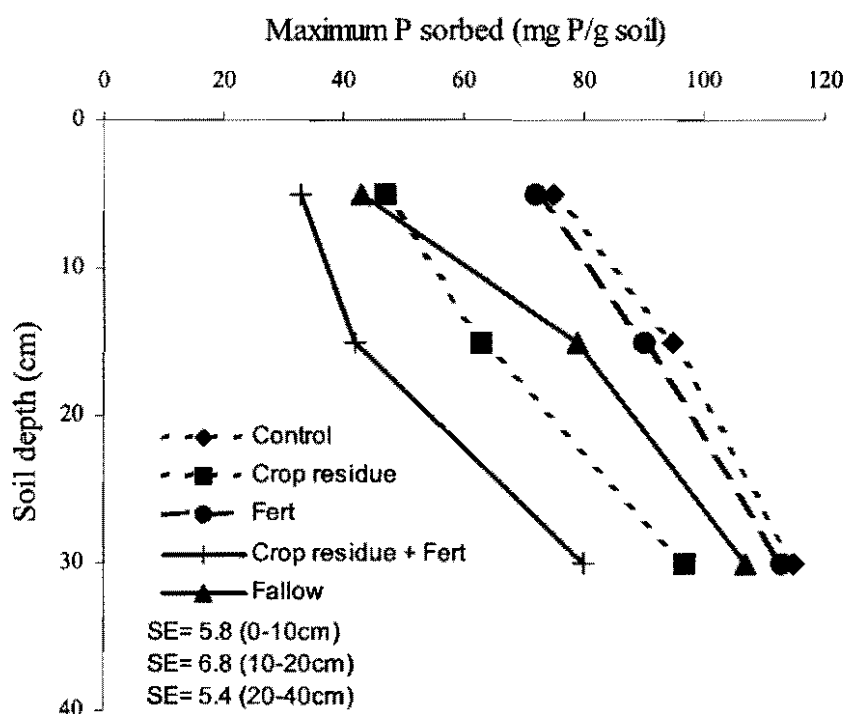
280 2. Calculated as (Yield Crop Residues + Fertilizer - Yield Control) / P applied

281 NA = not available

282 Source: Bationo et al. (1995)

283 The addition of manure and crop residue either alone or in combination with inorganic
 284 fertilizers frequently resulted in a substantial decrease in the soil's capacity to fix P. The
 285 maximum sorption of phosphorus calculated using the Langmuir Equation (Langmuir,

286 1918) decreased with the application of organic material (Figure 8). This may at least
 287 partly explain the demonstrated increase of P-fertilizer use efficiency with organic inputs.
 288 In laboratory experiments using the sandy Sahelian soils of West Africa, Kretschmar et
 289 al. (1991) found that the addition of crop residue resulted in an increased P availability
 290 which was attributed to the complexing of iron and aluminium by organic acids (Bationo
 291 et al., 1995).



292

293 **Figure 8.** Effect of soil amendments on maximum phosphorus sorbed in Sadore, Niger,
 294 1991 (Source: Bationo et al., 1995)

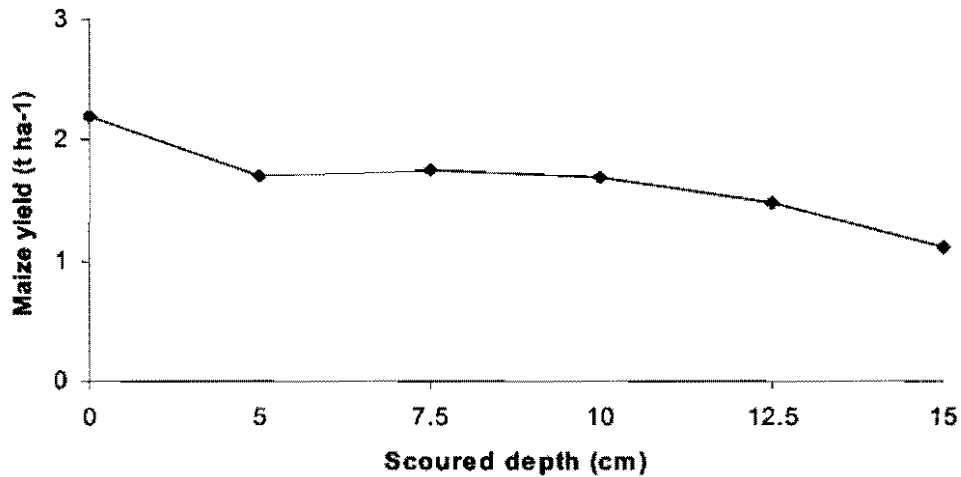
295 **5) Effect of soil and crop management on Soil Organic Carbon**

296 Soil organic carbon is lost through erosion, runoff and leaching (Roose and Barthes,
 297 2001). Erosion and runoff contribute a large portion of carbon losses and these are highly
 298 accelerated in cultivated land as compared to undisturbed forest or savanna (**Table 8**).
 299 Topsoil nutrients and organic carbon generally decrease with increasing erosion (Kaihura
 300 et al., 1998) with the amount of eroded carbon depending more on the erosion quantity
 301 than on the carbon content of the eroded sediments (Roose 1980). Roose and Barthes
 302 (2001) illustrated the importance of preserving the topsoil when by de-surfacing 5 cm of
 303 topsoil noted yield decline by about one-third in north Cameroon (**Figure 9**).

304 **Table 8.** Carbon losses (kg ha⁻¹ yr⁻¹) by erosion, runoff and leaching in the topsoil
 305 (30cm) in runoff plots at Adiopodoume, Korogho (Ivory Coast) and Saria
 306 (Burkina Faso)

C losses	Adiopodoume (2100 mm rainfall)		Korogho (1300 mm rainfall)		Saria (800 mm rainfall)	
	Sub- equatorial forest (undisturbed)	Cereal cultivation	Sudanian savanna (Undisturbed)	Cereal with fertilizers	Sudano- sahelian savanna (undisturbed)	Cereal cultivation
Erosion	13	1801	6	65	9	150
Runoff	1	65	2	18	1	5
Leaching	74	7	13	3	2	1
Total	88	1873	21	86	12	156

307 Adopted from Roose and Barthes (2001)



308

309 **Figure 9.** Effect of depth of soil mechanical desurfacing on maize grain yield at Mbissiri,
 310 North Cameroon (Source: Roose and Barthes, 2001)

311 The importance of soil textural (clay and silt) properties for the SOC content of soils was
 312 stressed repeatedly as clays are an important component in the direct stabilization of
 313 organic molecules and microorganisms (Amato and Ladd, 1992; Greenland and Nye,
 314 1959; Feller et al., 1992). Thus Feller et al. (1992) reported, that independent of climatic
 315 variations such as precipitation, temperature and duration of the dry season, SOC
 316 increased with the clay and silt contents but there was a poor relationship with the amount
 317 of rainfall. Therefore, small variations in topsoil texture at the field or watershed level
 318 could have large effects on SOC.

319 There is much evidence for a rapid decline of SOC levels with continuous cultivation of
 320 crops in West Africa (Bationo et al., 1995). For the sandy soils, average annual losses in
 321 SOC, often expressed by the k-value (calculated as the percentage of organic carbon lost
 322 per year), may be as high as 4.7%, whereas for sandy loam soils, reported losses seem

323 much lower at an average of 2%(Pieri, 1989; **Table 9**). Topsoil erosion may lead to
324 significant increases in annual SOC losses such as from 2% to 6.3% at the Centre de
325 Formation des Jeunes Agriculteurs (CFJA) in Burkina Faso (**Table 9**). However, such
326 declines are site-specific and heavily depend on management practices such as the choice
327 of the cropping system, soil tillage and the application of mineral and organic soil
328 amendments.

329 **Table 9.** Annual loss rates of soil organic carbon measured at selected research stations in
 330 the SSWA

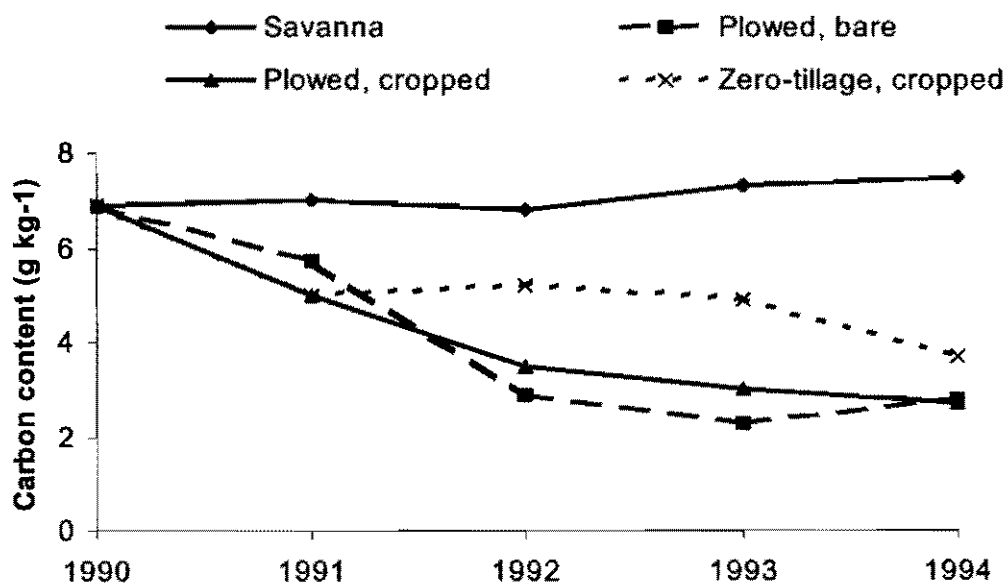
Place and Source	Dominant cultural succession	Observations	Clay + Silt (%) (0-0.2 m)	Annual loss rates of soil organic carbon (k)	
				Years of measurement	k (%)
Burkina Faso		With tillage			
Saria, INERA-IRAT	Sorghum monoculture	Without fertiliser	12	10	1.5
	Sorghum monoculture	Low fertilizser (lf)	12	10	1.9
	Sorghum monoculture	High fertiliser (hf)	12	10	2.6
	Sorghum monoculture	lf + crop residues	12	10	2.2
CFJA, INERA-IRCT	Cotton-cereals	Eroded watershed	19	15	6.3
Senegal		With tillage			
Bambey, ISRA-IRAT	Millet-groundnut	Without fertiliser	3	5	7.0
	Millet-groundnut	With fertiliser	3	5	4.3
	Millet-groundnut	Fertiliser + straw	3	5	6.0
Bambey, ISRA-IRAT	Millet monoculture	with PK fertiliser + tillage	4	3	4.6
Nioro-du-Rip, IRAT-ISRA	Cereal-leguminous	F0T0	11	17	3.8
	Cereal-leguminous	F0T2	11	17	5.2
	Cereal-leguminous	F2T0	11	17	3.2
	Cereal-leguminous	F2T2	11	17	3.9
	Cereal-leguminous	F1T1	11	17	4.7
Chad		With tillage, high fertility soil			
Bebedjia, IRCT-IRA	Cotton monoculture		11	20	2.8
	Cotton - cereals			20	2.4
	+ 2 years fallow			20	1.2
	+ 4 years fallow			20	0.5

331 F0 = no fertiliser, F1 = 200 kg ha⁻¹ of NPK fertiliser, F2 = 400 kg ha⁻¹ of NPK fertiliser + Taiba phosphate
 332 rock, T0 = manual tillage, T1 = light tillage, T2 = heavy tillage.

333 *Source: Pieri, (1989)*

334 Farming systems and cultural practices such as minimum tillage with crop residue can
 335 change erosion rate and SOC balance quite rapidly (Roose and Barthes, 2001; Pedro et
 336 al., 2001; Six et al., 2002). Accelerated mineralization following land clearing and
 337 continuous cropping has been reported to decrease SOC by up to 30% (Gregorich et al.,
 338 1998; Nandwa, 2001). Similarly, carbon losses by erosion from cropped land can be 4-20

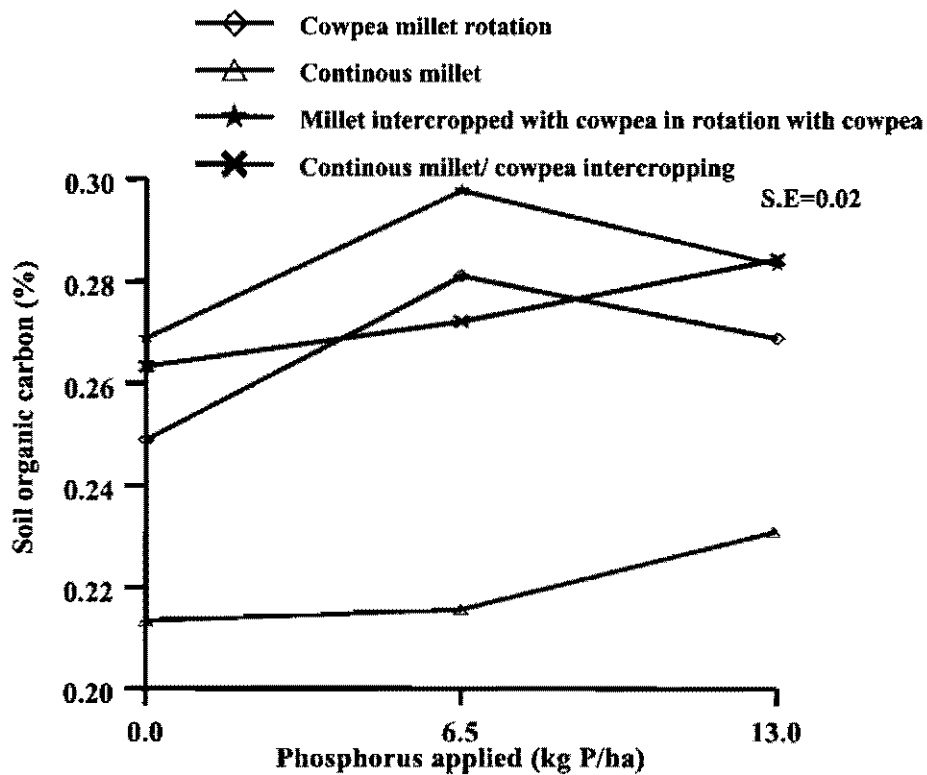
339 times higher than on natural sites (Roose and Barthes, 2001). In a study by Roose and
 340 Barthes (2001) in Cameroon, a sharp decline in SOC was observed in the top layer (0-10
 341 cm depth) of the conventional tillage due to accelerated mineralization (**Figure 10**).
 342 Under minimum tillage system, the decrease in SOC was slower because the topsoil was
 343 less disturbed. Pieri (1989) reported that without mineral fertilizer application, soil tillage
 344 increased annual rate of SOC losses from 3.8% (with manual tillage) to 4.7% and 5.2%
 345 following light and heavy tillage respectively.



346
 347 **Figure 10.** Evolution of carbon content in the 0-10 cm horizon, as affected by time and
 348 treatment in runoff plots of Mbissiri Station, Cameroon (Adapted from Roose
 349 and Barthes, 2001)

350 Rotations and intercropping systems have been reported by several authors to contribute
 351 to conservation of SOC. In Chad, cotton-cereal rotations reduced SOC losses from 2.8%
 352 in continuous cotton system to 2.4% in rotation systems (Pieri, 1989). Similarly, a

353 rotation trial at Sadoré in the Sahel revealed significant effects of crop rotation on SOC
 354 contents. After five years, SOC levels were 2.8 g kg⁻¹ in millet/cowpea intercrop plots
 355 that were rotated with pure cowpea compared to continuous millet plots with 2.2 g SOC
 356 kg⁻¹ (Bationo and Buerkert, 2001). The higher SOC level in the cowpea system was at
 357 least partly due to the falling of leaves from the legume crop. **Figure 11** shows the effect
 358 of cowpea-millet rotation, millet-cowpea-intercrop rotated with cowpea and continuous
 359 millet cowpea intercrop on SOC.



360

361 **Figure 11.** Effect of cropping system and phosphorus on SOC in Sadoré, Niger in 1995

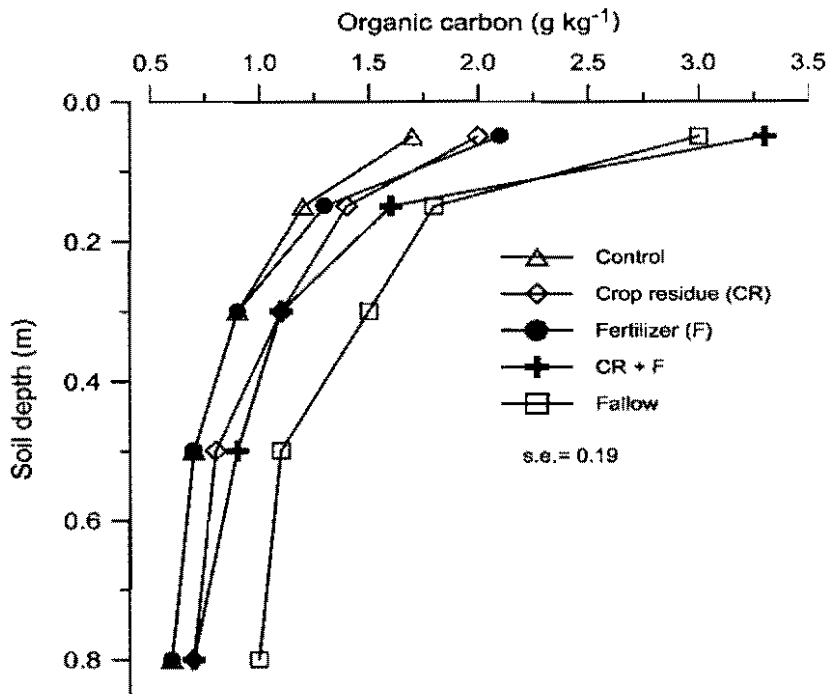
362 Mulching decreases soil temperature, maintains favorable soil structure and infiltration
 363 rate, and enhance microbial and mesofaunal activities (Lal, 1975; Boli et al., 1993; Roose

364 and Barthes, 2001). Mulches also contribute to carbon stock through their mineralization
365 and the effect of reduced erosion (Nandwa, 2001).

366 Lone application of mineral fertilizer can cause decline in soil organic carbon. Pichot et al
367 (1981) reported from a ferruginous soil in Burkina Faso that with mineral fertilizer
368 application, 25-50% of the indigenous organic matter disappeared during the first two
369 years of cultivation. Bache and Heathcote (1969), Mokwunye (1981) and Pichot et al.
370 (1981) observed that continuous cultivation using mineral fertilizers increased nutrient
371 leaching, lowered the base saturation and aggravated soil acidification. Also
372 exchangeable aluminium was increased and crop yield declined.

373 Application of organic material such as green manures, crop residues, compost or animal
374 manure can counteract the negative effects of mineral fertilizers (de Ridder and Van
375 Keulen, 1990). This led Pieri (1986) to conclude that soil fertility in intensive arable
376 farming in WASAT can only be maintained through efficient recycling of organic
377 material in combination with rotations of N₂-fixing leguminous species and chemical
378 fertilizers.

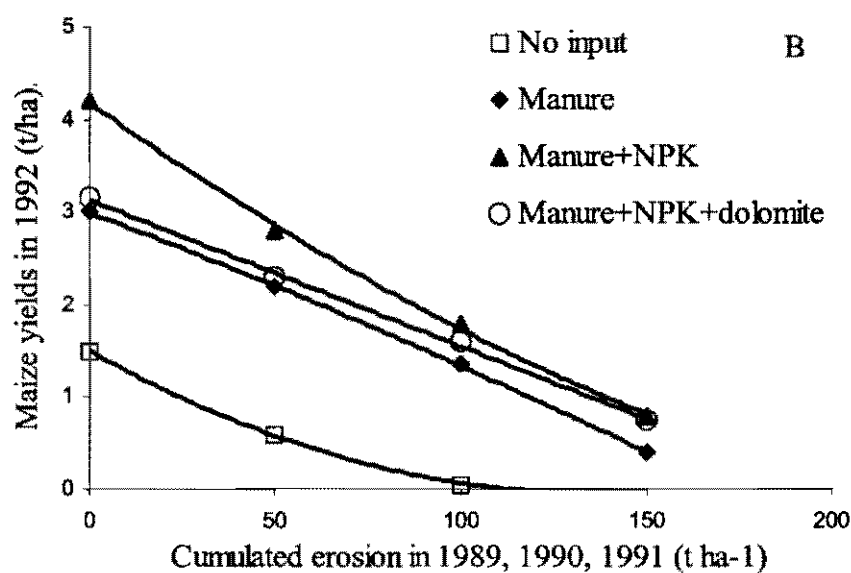
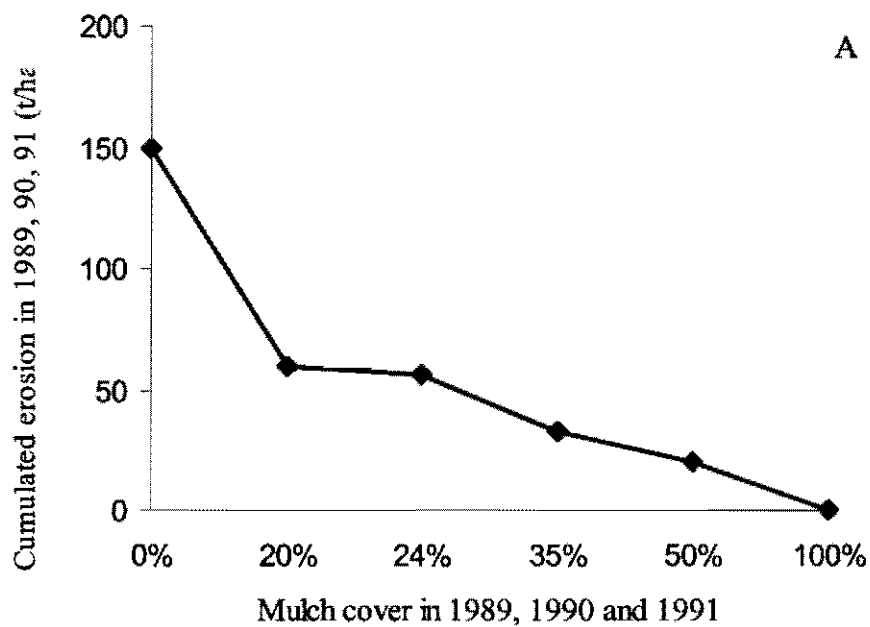
379 In a long-term crop residue management trial in the Saheian Zone, Bationo and Buerkert
380 (2001) found that levels of SOC were 1.7 g kg⁻¹ and 3.3 g kg⁻¹ respectively at 0.1m for 2t
381 ha⁻¹ and 4 t ha⁻¹ of mulching with crop residue applied as compared to unmulched plot
382 **(Figure 12)**



383

384 **Figure 12.** Soil organic carbon (SOC) as affected by soil depth and management
 385 practices. Sadoré, Niger, rainy season, 1996.

386 Another example from Central Africa reported the importance of mulching on soil losses
 387 and subsequent productivity restoration (**Figure 13**). Total erosion during three years
 388 decreases from 154t ha⁻¹ on bare fallow to 0.15t ha⁻¹ on entirely mulched plots. The effect
 389 of erosion on subsequent land productivity reported maize yield of over 4t ha⁻¹ on non-
 390 eroded plots against yields of less than 1t ha⁻¹ on eroded plots (soil loss of 150t ha⁻¹) even
 391 with the application of mineral and organic plant nutrients (Rushirumuhirwa 1997,
 392 Rushirumuhirwa and Roose 1998).



393

394 **Figure 13:** Influence of mulch cover (linked to banana plantation density) on erosion
 395 (13a), and effect of this previous erosion on grain yield of the next maize crop
 396 (13b), in runoff plots (8% slope) of Mashitsi station (Burundi). During maize

397 crop, manure input was 20 t ha⁻¹, fertilizer level was 60N, 40P, 60K and
 398 dolomite, 500kg ha⁻¹. Source: Roose and Barthes, 2001.

399 The data in **Table 10** shows that manure collected from stables and applied alone
 400 produces about 20 to 60 kg N ha⁻¹ in cereal grain and 70 to 178 kg of kg N ha⁻¹ in stover
 401 per tonne of manure.

402 **Table 10:** Results of manuring experiments at three sites in semi-arid West Africa

403 **Panel A: Manure only**

Location	Amount of manure applied (t ha ⁻¹)	Crop	Crop response ¹ (kg of DM t ⁻¹ manure)		Reference
			Grain	Stover	
M□Pesoba, Mali	10	Sorghum	35 ²	n.s.	1
Saria, Burkina Faso	10	Sorghum	58	n.s.	2
Sadore, Niger 1987	5	Pearl millet	38	178	3
	20	Pearl millet	34	106	3

404 **Panel B: Manure with inorganic fertilizer**

Location	Amount of		Crop	Crop response ¹ (kg of DM/t manure ⁰)	
	Manure (t/ha)	Fertiliser (kg/ha)		Grain	Stover
M□Pesoba, Mali	5	NPK: 8-20-	Sorghum	90 ³	n.s.

Saria, Burkina Faso	10	Urea N: 60	Sorghum	80	n.s.
Sadore, Niger 1987	5	SSP P: 8.7	Pearl millet	82	192
Sadore, Niger 1987	20	SSP P: 17.5	Pearl millet	32	84

405 1. Responses were calculated at the reported treatment means for crop yields as:

406 (treatment yield - control yield)/quantity of manure applied.

407 2. Response of sorghum planted in the second year of a 4-year rotation involving cotton-
408 sorghum-groundnut-sorghum. Manure was applied in the first year.

409 3. Estimated from visual interpolation of graph

410 n.s. implies not specified

411

412 References: 1. Pieri (1989); 2. Pieri (1986); 3. Baidu-Forson and Bationo (1992)

413 Source: Williams et al. 1995

414

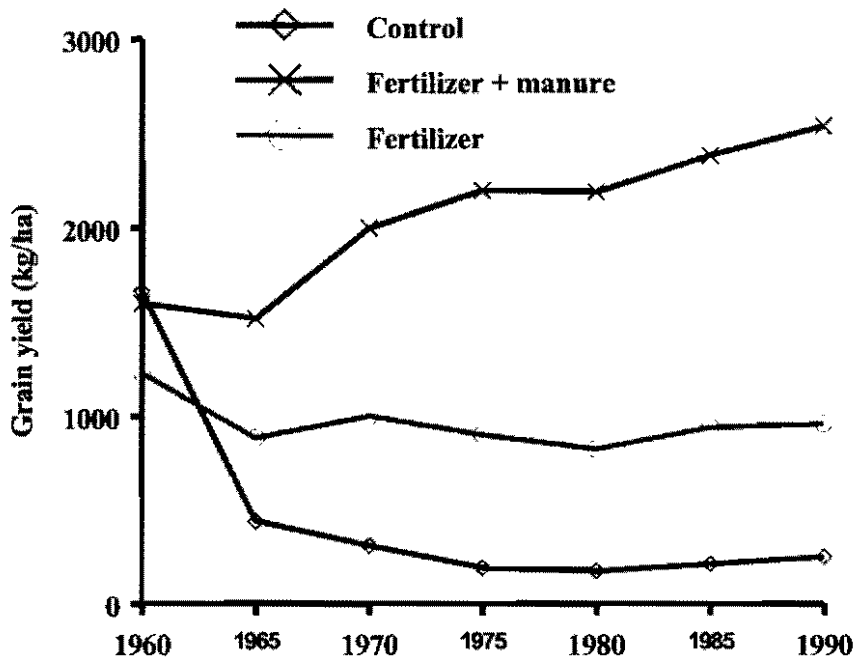
415 The data in **Table 11** indicates that the application of 3 t ha⁻¹ of manure plus urine
416 produced grain and total bio-mass that were three to four times as high compared to when
417 only manure was applied and crop response to sheep manure was greater than to cattle
418 manure. Research studies indicate that approximately 80-95% of the N and P consumed
419 by livestock is excreted. Whereas N is voided in both urine and faeces, most P is voided
420 in faeces (ARC, 1980; Termouth, 1989).

421 **Table 11:** Effect of cattle and sheep dung and urine on pearl millet grain and total above-
 422 ground biomass, Sadore, Niger 1991

Type of manure	Dung application rate kg ha ⁻¹	With urine		Without urine	
		Grain yield (kg ha ⁻¹)	Total biomass (kg ha ⁻¹)	Grain yield (kg ha ⁻¹)	Total biomass (kg ha ⁻¹)
Cattle	0	-	-	80	940
	2990	580	4170	320	2170
	6080	1150	7030	470	3850
	7360	1710	9290	560	3770
	s.e.m.	175	812	109	496
Sheep	0	-	-	80	940
	2010	340	2070	410	2440
	3530	1090	6100	380	2160
	6400	1170	6650	480	2970
	s.e.m	154	931	78	339

423 Adapted from Powell et al. 1998

424 One important conclusion that emerged from the long-term experiments is that
 425 application of mineral fertilizers is an effective technique for increasing crop yields in the
 426 Sudanian zone of West Africa. However, in the long run the use of mineral fertilizers
 427 alone will decrease crop yields but sustainable and higher production is obtained when
 428 inorganic fertilizers are combined with manure (**Figure 14**).

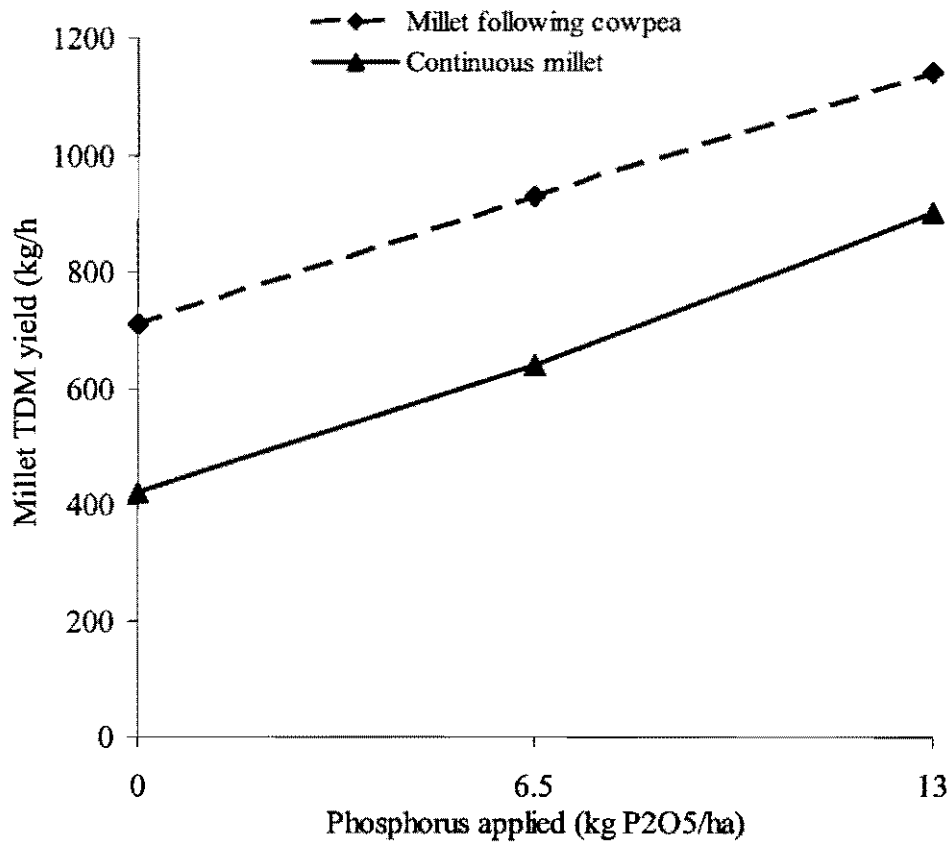


429

430 **Figure 14:** Sorghum grain yield as affected by mineral and organic fertilizers over time.

431 Source: Sedogo (1993)

432 As already indicated, SOC is significantly higher in rotation or intercropping systems of
 433 pearl millet and cowpea and this is one of the reason of higher productivity of millet in
 434 the rotation than in monoculture system (**Figure 15**)



435

436 **Figure 15:** Effect of P fertilization and rotation on millet total dry matter yield

437 **Figure 16** gives a schematic representation of the different uses of crop residues.

438 Traditionally, many farmers burn whatever is left of their crop residue once their needs

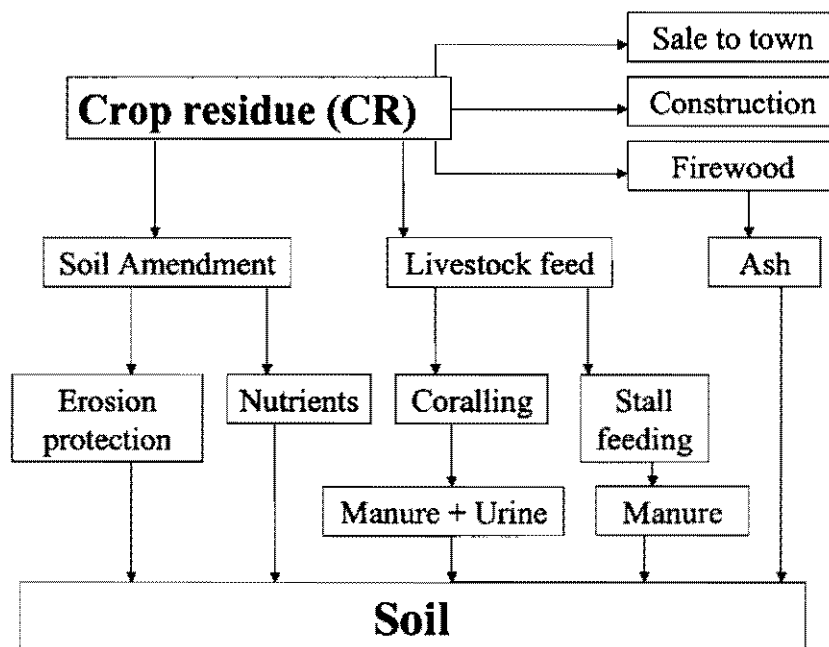
439 for fuel, animal feed, or housing and fencing material have been fulfilled. In West Africa,

440 grazing animals remove more biomass and nutrients from cropland than they return in the

441 form of manure. Therefore, Breman and Traore (1986) concluded that a sustainable

442 nutrient supply in the southern Sahel based on a net transfer of nutrients from rangelands

443 to cropland required between 4 and 40 ha of rangeland per hectare of cropland.



Source: Bationo et al., 1995

444

445 **Figure 16:** The competing uses of crop residues in the West Africa Semi Arid Tropics

446 Availability of organic inputs in sufficient quantities and quality is one of the main
 447 challenges facing farmers and researchers today. In an inventory of crop residue
 448 availability in the Sudanian zone of central Burkina Faso, Sedga (1991) concluded that
 449 the production of cereal straw can meet the currently recommended optimum level of 5t
 450 ha⁻¹ every two years. However, the competition with other uses was not accounted for in
 451 this study. Lompo (1983) found in that zone upto 90% of crop residue is burned for
 452 cooking. This practise results in considerable loss of carbon and nutrients such as
 453 nitrogen and sulfur. Charreau and Poulain (1964) reported that 20 to 40kg N ha⁻¹ and 5 to
 454 10 kg S ha⁻¹ are lost by burning crop residues. Other negative effects might be temporal
 455 changes in the population of micro organisms in the upper soil layers, particularly
 456 rhizobia , by the intense heat (Charreau and Nicou, 1971). Increasing the availability of
 457 crop residue to maintain soil fertility in West Africa will require enhanced fuel

458 production to which agroforestry research might make a contribution by screening locally
459 adapted fast-growing woody species.

460 For the Sahelian zone, field experiments in millet showed that from a plant nutritional
461 standpoint the optimum level of crop residue to be applied to the soil as mulch may be as
462 high as 2t ha⁻¹ (Rebafka et al., 1994). However, McIntire and Fussell (1986) reported that
463 on fields of unfertilized local cultivars, grain yield averaged only 236 kg ha⁻¹ and mean
464 residue yields barely reached 1300 kg ha⁻¹. These results imply that unless stover
465 production is increased through application of fertilizers and or manure it is unlikely that
466 the recommended levels of crop residue could be available for use as mulch.

467 In village level studies on crop residue along a north-south transect in three different
468 agro-ecological zones of Niger, surveys were conducted to assess farm-level stover
469 production, household requirements and residual stover remaining on-farm. The results of
470 these surveys showed that the average amounts of stover removed from the field by a
471 household represented only between 2 to 3.5% of the mean stover production (ICRISAT,
472 1993). At the onset of the rains the residual stover on-farm was only between 21 and 39%
473 of the mean stover production at harvest time. Even if no data have been collected on the
474 amount of crop residue lost by microbial decomposition and termites, cattle grazing is
475 likely to be responsible for most of the disappearance of crop residues. Similar losses
476 were reported by Powell (1985) who found that up to 49% of sorghum and 57% of millet
477 stover disappearance on the humid zone of Nigeria was due to livestock grazing.
478 Sandford (1989) reported that in the mixed farming systems, cattle derive up to 45% of
479 their total annual intake from crop residues and up to 80% during periods of fodder

480 shortage. Up to 50% of the total amount of crop residue and up to 100% of the leaves are
481 eaten by livestock (van Raay and de Leeuw, 1971). Most of the nutrients are voided in
482 the animal excreta but when the animals are not stabled, nutrients contained in the
483 droppings cannot be effectively utilized in the arable areas (Balasubramanian and Nnadi,
484 1980).

485 In an on-farm crop residue availability study, Bationo et al. (1991) showed that the use of
486 fertilizers increased stover yields under on-farm conditions. Despite many competing
487 uses of crop residue as already mentioned (Figure 16) the increased production led to
488 significantly more mulch in the subsequent rainy season.

489 The availability of manure for sustainable crop production has been addressed by several
490 scientists. De Leeuw et al. (1995) reported that with the present livestock systems in West
491 Africa, the potential annual transfer of nutrient from manure will be 2.5kg N and 0.6 kg P
492 per hectare of cropland. Although the manure rates applied are between 5 and 20t ha⁻¹ in
493 most of the on-station experiments, quantities used by farmers are very low and ranged
494 from 1300 to 3800 kg ha⁻¹ (Williams et al., 1995).

495 ***6) Future research challenge with emphasis on organic matter quantity and quality***

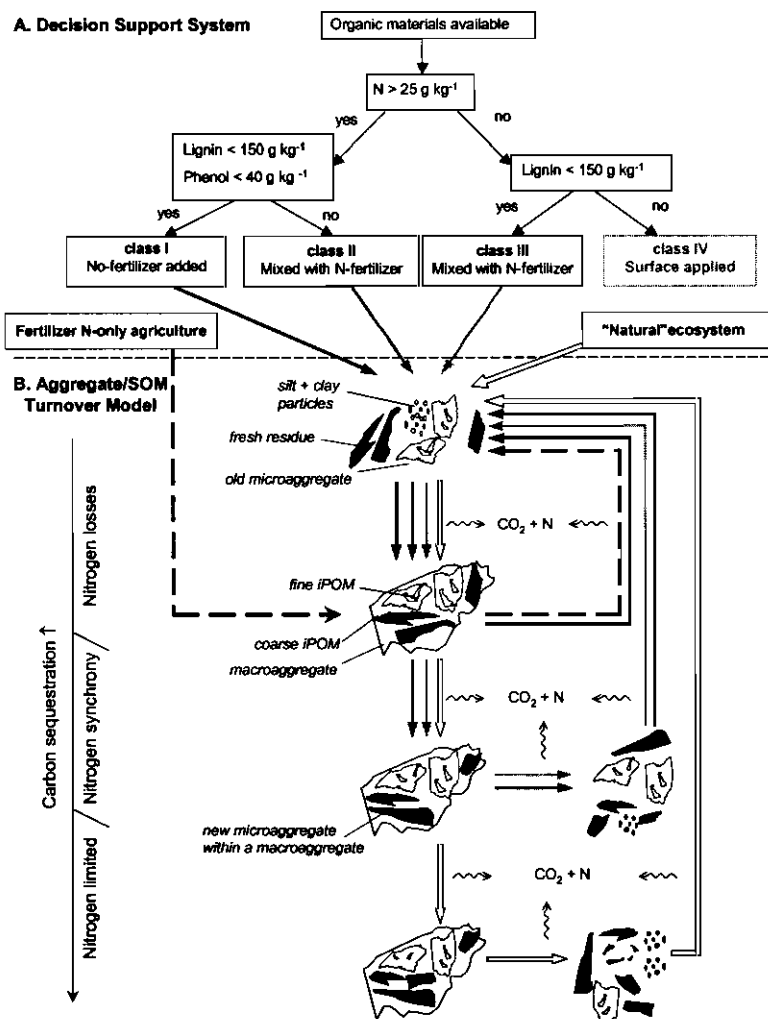
496 The complementarity of livestock and crop production suggests the need for research on
497 possibilities to increase nutrient use efficiency for higher crop residue production and to
498 improve the production of alternative feed supplies. The aim of such research should be
499 to increase both fodder quantity and quality thus preserving more crop residue for soil
500 application. There is need to increase crop biomass at farm level and future research

501 should focus on improvement of nutrient use efficiency in order to increase crop biomass.
502 Future research should alleviate socio-economic constraints in order to increase the
503 legume component in the cropping systems. This will produce higher quality fodder for
504 the livestock and also increase biomass at farm-level.

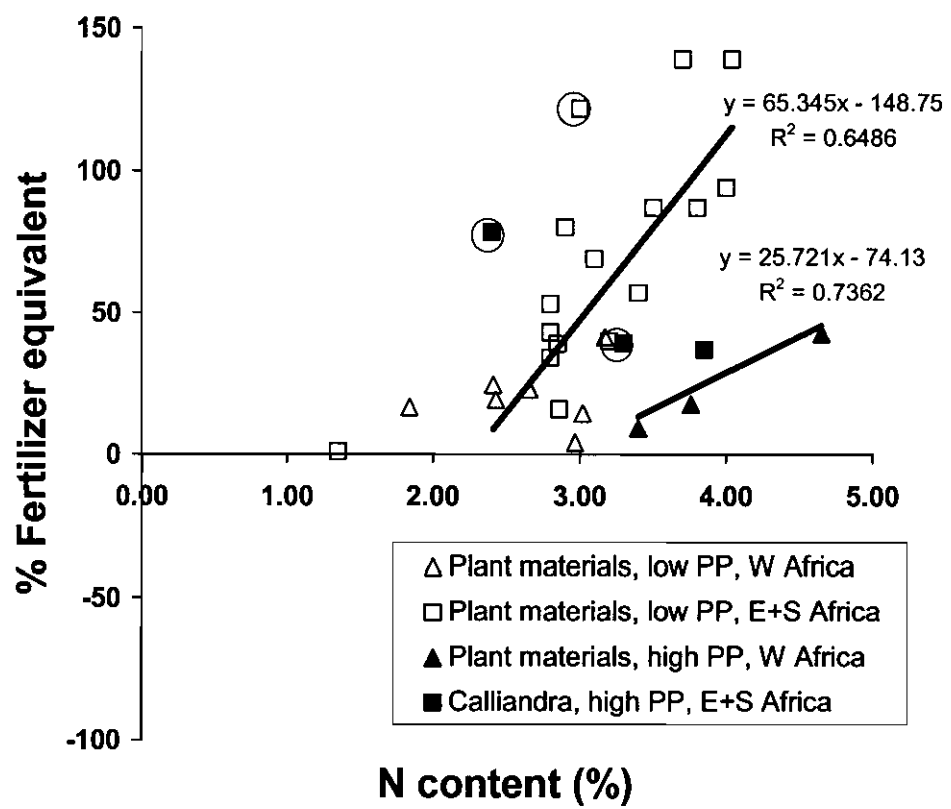
505 In the decision support system for organic matter management, recommendation for
506 appropriate use of organic material was made based on their resource quality, expressed
507 as a function of N, polyphenol and lignin content (**Figure 17a**). The fertilizer equivalency
508 of the different organic material can be predicted by the N content and the polyphenol
509 contents (**Figure 18**). Applications of high quality organic materials enhance quick
510 transformation into labile SOM fractions and improve nutrient supply and availability
511 (Vanlauwe, 2004; Nandwa, 2001). The impact of organic resource quality on SOC is less
512 clear. Roose and Barthes (2001) noted that the application of easily mineralizable manure
513 was not sufficient to increase SOC levels. Low quality organic resources, which show
514 limited increases in crop growth, contain substantial amounts of soluble polyphenols and
515 lignins that may affect the longer-term decomposition dynamics and contribute to build
516 up of SOC (Vanlauwe, 2004) necessary in the improvement of soil structure, aggregate
517 stability (Six et al., 2002) and soil water buffering (Nandwa, 2001). Additionally, these
518 resources have a positive impact on the environmental service functions of the soil
519 resource (Vanlauwe, 2004).

520 Future research will therefore need to focus more on whether the organic resource quality
521 concept is also useful for predicting different degrees of stabilization of applied organic C
522 in one or more of the organic matter pools. A recent hypothesis being tested is focusing

523 on the linkage between resource quality, aggregate turnover, N use efficiency and C
 524 cycling across soil textures and climates (Figure 17 a and b). In natural ecosystems,
 525 nutrient limitation induces a slow aggregate turnover, which leads to a C sequestration
 526 and reinforces the nutrient limited environment. In intensively managed agroecosystems,
 527 aggregate turnover is fast due to high N availability and disturbance leading to high N
 528 and carbon losses. The combined use of organic resources of intermediate quality and
 529 mineral resources under this hypothesis may result in an optimal balance between N and
 530 C stabilization versus N availability for plants.



532 **Figure 17.** The decision support system for organic matter management and SOM
533 turnover (Source: Palm et al., 2001; Six et al., 2000)



534

535 **Figure 18.** Fertilizer equivalency for different organic materials

536 (Source: Vanlauwe et al., 2002)

537

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