

2 **Effects of organic and inorganic fertilization on soil bacterial and fungal microbial**
3 **diversity in the Kabete Long-term trial, Kenya**

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25 **ABSTRACT**

26 The effects of crop manure and inorganic fertilizers on composition of microbial communities of
27 central high land soils of Kenya are poorly known. For this reason, we have carried out on a thirty-two
28 year old long-term trial in Kabete, Kenya. These soils were treated with organic (maize stover (MS) at
29 10 t ha^{-1} , farmyard manure (FYM) at 10 t ha^{-1}) and inorganic fertilizers 120 kg N , 52.8 kg P (N_2P_2),
30 $\text{N}_2\text{P}_2+\text{MS}$, $\text{N}_2\text{P}_2+\text{FYM}$, a control (C) and a fallow (F) for over 30 years. We examined 16S rRNA gene
31 and 28S rRNA gene fingerprints of bacterial and fungal diversity, by PCR amplification respectively
32 and denaturing gradient gel electrophoresis (PCR-DGGE) separation. The PCR bacterial community
33 structure and diversity was negatively affected by N_2P_2 and was more closely related to the bacterial
34 structure in the soils without any addition (control) than that of soils with a combination of inorganic
35 and organic or inorganic fertilizers alone. The effect on the composition fungal diversity by N_2P_2 was
36 different than the effect on the composition of bacterial diversity since the fungal diversity was similar
37 to that of the $\text{N}_2\text{P}_2+\text{FYM}$ and $\text{N}_2\text{P}_2+\text{MS}$ treated. However, soils treated with organic inputs clustered
38 away from soils amended with inorganic inputs. Organic inputs had a positive effect on both bacterial
39 and fungal diversity with or without chemical fertilizers. Results from this study suggested that total
40 diversity of bacterial and fungal communities was closely related to agro-ecosystem management
41 practices and may partially explain the yield differences observed between the different treatments.

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43 Key words: Organic and Inorganic amendments, Microbial diversity.

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45

46 **Introduction**

47

48 In large parts of Kenya, food crop yields are very poor due to declining soil fertility as a result of
49 continuous cropping with little or no replenishment of nutrients removed through either crop harvests
50 or other losses such as leaching and soil erosion (Kibunja 2007). With the liberalization of trade and
51 introduction of Structural Adjustment Programs (SAP), unfavorable crop fertilizer prices and financial
52 constraints have greatly contributed to the low level of fertilizer use in many small-holder' farms in
53 the country (Waswa et al. 2007). The Eastern and Central highlands smallholder farmers in particular,
54 have been experiencing declining soil fertility and crop productivity. This has compelled researchers
55 to place a high priority on increasing agricultural productivity and alleviation of poverty among small-
56 holder farmers. Attainment of this goal calls for an Integrated Soil Fertility Management (ISFM), with
57 management practices that necessarily include the use of fertilizer, organic inputs, and improved
58 germplasm combined with the knowledge on how to adapt these practices to local conditions, aiming
59 at maximizing agronomic use efficiency of the applied nutrients and improving crop productivity
60 (Vanlauwe et al. 2010). In response to these problems, researchers in collaboration with farmers have
61 set up trials on various soils management techniques for combating soil nutrient depletion.
62 Technologies being actively and vigorously promoted to farmers in the region include: the combined
63 use of soil amendments, organic materials, and mineral fertilizers to replenish the soil nutrients and
64 improve the efficiency and cost-effectiveness of external inputs. These technologies have been found
65 to be both technically feasible and socially acceptable (Sanchez and Jama 2000; Jama et al. 1999). The
66 emerging dilemma with this approach however, is that the impacts of these resources on soil biota
67 which are key functional groups influencing agricultural production and ecosystem services, are not
68 yet well documented.

69 Intensification of agriculture has focused on the use of chemical and mechanical inputs, often
70 at the expense of biologically-mediated processes. Soil biota is indispensable for key soil function
71 such as decomposition of soil organic matter, nutrient cycling and formation of soil aggregates
72 (Sparling 1997). The total mass of living microorganisms therefore has a central role as a source, sink
73 and regulator of the transformation energy and nutrients in the soil. Both organic and inorganic
74 sources of fertilizer have residue effects in the field. These effects are a vital component of
75 sustainability because they smooth season-to-season variations in soil fertility and crop productivity,
76 but they are difficult to assess quantitatively. Therefore, it is advantageous to undertake well-
77 characterized medium-to-long term experiments rather than single-season trials, and to detail the
78 interactions rather than averaging the responses over different seasons and environments (Tandon and
79 Kanwar 1984). Long-term implies that primary objectives, treatment and management are not changed
80 during the period under consideration, often regarded as at least 10 years (Laryea et al. 1995). An
81 experiment was established in 1976 at Kabete and the primary objective of the trial was to identify
82 appropriate methods for maintaining and improving the productivity of the soil through repeated use
83 of inorganic fertilizers (in particular nitrogen and phosphorus), farmyard manure and crop residues
84 under continuous cropping following the husbandry practices of small-scale farmers. The challenge
85 now remains to provide some understanding on the impact of these resources on soil microbial
86 communities. The work described here was carried out in the long-term field experiment at the
87 National Agricultural Research Laboratories (NARL-KARI) Kabete, near Nairobi, Kenya, which
88 provides a unique resource to investigate of the impact of long-term fertilization of organic and inorganic
89 resources on the soil bacterial and fungal communities.

90

91 **Materials and Methods**

92

93 **The study site description**

94 The study was superimposed on the on-going long-term field experiment, at Kenya Agricultural
95 Research Institute at the National Agricultural Research Laboratories (NARLs) station at Kabete,
96 located at 36° 41'E and 01°15'S and at an altitude of 1737 m above sea level (Kibunja et al. 2010).
97 These soils are mainly humic Nitisols (FAO, 1990) that are deep and well weathered. The mean
98 annual rainfall is about 980 mm is divided into two distinct annual rainy seasons; the long rains (LR)
99 between mid-March and June, and the short rains (SR) between mid-October and December. The
100 average annual maximum and minimum temperature is 23.8 °C and 12.6 °C, respectively.

101

102 **Experimental design and choice of treatments**

103 The field experiment was established in 1976 to investigate the effect of continuous application of
104 farmyard manure, crop residues and nitrogen (N) and phosphorus (P) fertilizers in a maize-bean
105 rotation. The experiment was established as a randomized complete block design (RCBD) with
106 eighteen treatments replicated four times (4 blocks). For this study, only seven treatments were
107 monitored for a period of one year from March 2007 to February 2008 and included one maize and
108 bean-growing season. Each plot is a rectangle with an area of 30 m² (4 x 7.5 m²). The selection of the
109 seven treatments (Table 1) was based on the crop yields data from the KARI yields data-base. The
110 averages of the soils characteristics are as follows: Organic Matter (3.67%), total N (0.16%), total C
111 (2.13%) and available P (163 mg of P kg⁻¹).

112

113 **Experimental design for bacterial and fungal communities' diversity**

114 Soil samples were collected six weeks after planting following a 'W' design, across the plots receiving
115 treatments of interest (Table 1), by pushing a sterile soil auger at a depth of (0-10 cm). Six soil cores
116 per plot were taken and well mixed to make a composite sample of 20 g. This was repeated for the 4
117 blocks. The soils were put in zip lock plastic bags and transported in cooler boxes to the laboratory
118 where they were frozen at -80° C degrees and analyzed within the three weeks.

119

120 **DNA extraction, PCR amplification and DGGE**

121 Total genomic DNA was extracted from all soil samples by the direct lysis extraction procedure
122 (Martin-Laurent et al. 2001). Soil (0.25 g) was treated using 0.5 g glass beads (106µm diameter,
123 Biospec Products) and 1ml lysis buffer (100mM Tris-HCl (pH 8.0), 100mM NaCl, 1% (w/v)
124 polyvinylpolypyrrolidone (PVPP), and 2% (w/v) sodium dodecyl sulfate (SDS)) in a 2 ml Eppendorf
125 tube. The Eppendorff tubes were homogenized for 30 seconds at 1600 rpm using a Biospec Mini-
126 Bead- Beater cell disruptor, centrifuged at 14,000 g for 1 min at 4°C. The supernatant was collected,
127 1/10 volume of 5M sodium acetate added and incubated on ice for ten minutes, then centrifuged at
128 14,000 g for 5 minutes. The DNA was precipitated with one volume of ice cold isopropanol overnight.
129 The DNA was washed with 70% ethanol and dissolved in double distilled water.

130 PCR amplifications were performed using the forward primer 338f with a GC clamp and
131 reverse primer 518r (Øvreås and Torsvik 1998). The total reaction mixture (25 µl) contained 2 µl of
132 pure total DNA extract, 1X freeze dried bead (Ready-to-Go PCR beads, Pharmacia Biotech)
133 containing 1.5 U of Taq polymerase, 10mM Tris-HCl, (pH 9 at RT), 50 mM KCl, 1.5 mM MgCl₂, 200
134 µM of each dNTP and 1.0 µM of each primer and sterile distilled water. Two replicates were
135 performed for each sample. A Primus 96^{plus} PCR system (MWG AG BIOTECH) thermal cycler was
136 used for

137 PCR amplification with 2 min at 94°C followed by 30 cycles of 30 sec at 94°C, 30 sec at 55°C
138 and 1 min at 72°C. The first 20 cycles had an annealing temperature of 65°C, which decreased every
139 cycle until a touchdown at 55°C. The primer extension was carried out at 72°C for 15 min. The PCR
140 product (approximately 180 bp) was checked on a 1% agarose gel stained with ethidium bromide.
141 PCR amplifications were performed using the forward primer 662f with a GC clamp and reverse
142 primer 314r as described by Sigler et al. (2002).

143 PCR amplifications were carried out in 25- μ l reaction volumes with a Primus 96^{plus} PCR
144 system (MWG AG BIOTECH). The reaction mixture was slightly modified by reducing the quantity
145 of template DNA from 2.0 μ l to 1.0 μ l. Cycling conditions were as follows: initial denaturation at
146 95°C for 10 min followed by 49 cycles of denaturation at 95°C for 1 min, annealing at 50°C for 1 min,
147 and extension at 72°C for 2 min followed by a final extension phase at 72°C for 10 min. The PCR
148 product (approximately 348 bp) was checked for size on a 1% agarose gel stained with ethidium
149 bromide.

150 PCR products were separated using DGGE with a D-Code Universal Mutation system (Bio-
151 Rad Laboratories). The PCR products were loaded on 8% (w/v) polyacrylamide (acrylamide-
152 bisacrylamide [37.5:1]) gels with a denaturing gradient of 40%-60% (100% denaturant contains 7 M
153 urea and 40% formamide). The gels were run for 16 hours at 75V in Tris-acetate-EDTA buffer (1X) at
154 60°C with a pre-run of 10 min at 20V. After migration, the DGGE profiles were stained with EtBr for
155 20 min, destained for 45 min and then scanned with Gel DocTM XR (BIO-RAD Laboratories). The gel
156 images were captured using Quantity-One software version 4.6.0 (BIO-RAD Laboratories).

157 The DGGE patterns were transformed to graphs by the Bio-Rad Quantity OneTM software:
158 where each resolved band of the gel lanes was converted to a trait whose intensity is related to the
159 amount of the corresponding DNA fragment. Total Lab 120 version 2006 software (Nonlinear

160 Dynamics Ltd) was then used to calculate the percentage of similarity among lanes taking into account
161 the migration distance and the relative intensity of all bands. The DGGE profiles were compared and
162 unweighted-pair group method cluster analysis used to produce the dendograms by using the total lab
163 120 version 2006 software..

164

165 **Diversity measurements and Data analysis**

166 The Shannon-Weaver indices (H') Shannon and Weaver (1963) were calculated using the following
167 equation:

$$168 \quad H' = -\sum p_i \log p_i$$

169 Where $p_i = n/\sum n$

170 Where: n is the volume of a single band per lane and $\sum n$ is the total volume of all bands per lane.

171 The Shannon indexes were used to perform ANOVA analysis using GENSTAT Release 7.2 (Lawes
172 Agricultural Trust 2007).

173

174 **Results**

175 Genetic finger printing by DGGE of eubacteria 16S rRNA amplified fragment showed few dominating
176 bands and a number of faint unresolved bands (Fig 1a), compared to the PCR amplification of fungal
177 28S rDNA gene where most of the bands were distinct (strong) and well resolved (Fig 1b). However
178 the number of bands was higher in the eubacteria 16S rRNA gene DGGE profiles than in the fungal
179 28S rDNA gene DGGE profiles an indication of high bacteria numbers compared to the fungal
180 populations in the various treatments. Cluster analysis of PCR-DGGE patterns generally distinguished
181 the fungal and bacterial communities in two major clusters (Fig 2). In the bacteria community, cluster
182 analysis showed great similarity (65 %) with two major clusters, those with mineral fertilizers only,

183 fallow and control and the fallow in one cluster and those with inorganic and with or without inputs in
184 another cluster (fig 2a), while for the fungal communities, a moderately lower percentage (55 %) with
185 two major clusters was also observed, organically treated soils clustered away from those with
186 combined inorganic and organic inputs (Fig 2b). Within the group of the combined inorganically and
187 organically as well as that of inorganic treatments, the control and fallow treatments further grouped
188 into a smaller cluster indicating different community structures in the control and the fallow land.
189 Shannon indices indicated a higher total diversity of fungal and bacterial communities in the
190 N_2P_2 +FYM treatments. Lowest diversity in the bacterial community was in the N_2P_2 treated soils (Fig
191 3) while for the fungal communities the lowest diversity was observed in the fallow and control
192 treatments.

193 Statistical analysis of Shannon indices indicated that there were no significant differences ($p \leq$
194 0.05) between N_2P_2 +R, N_2P_2 +FYM and fallow treatments for bacterial communities (Table 2). A
195 significant difference was however noted between these treatments and N_2P_2 , Control, MS and FYM
196 (Table 2). For the fungal communities no significant differences ($p \leq 0.05$) were found between
197 N_2P_2 +MS and N_2P_2 +FYM but significant differences were noted between fallow, control, M, FYM
198 and N_2P_2 .

199

200 **Discussion**

201 While the significance of the microbial community shifts following soil management practices
202 remains to be recognized in the tropical regions, the results of this study indicate that the long-term
203 combination of organic and inorganic amendments enriches the soil bacterial and fungal community
204 and promotes diversity. Zhong et al (2010) showed how long-term fertilization of organic manure
205 (with or without NPK application) led to C utilization pattern shifts and increased soil microbial

206 functional diversity. Previous work in the same site showed that use of manure, NP fertilizers and crop
207 residue retention caused least organic C loss from the soil and raised the total soil N significantly
208 compared to the control (Kapkiyai et al.1998). Allison et al (2007) found that N fertilization alone
209 reduced fungal taxonomic richness and altered community structures, N deposition reducing
210 significantly total fungal diversity in soils (mainly by reducing decomposer fungi). This has been
211 confirmed by our results. Maly et al (2009) showed that long-term mineral fertilization increases the
212 proportion of *r*-strategists in soil. The authors assume that these results are probably due to increases
213 in available P and rhizodepositions. Meanwhile Kibunja et al. (2010) found continuous application of
214 chemical fertilizers in this site led to net loss of SOC and a drop in soil pH explaining the decline in
215 bacterial communities in this treatment. Our results also showed that bacterial community population
216 was generally higher than that of the fungal community in all treatments. Zhong et al (2010) suggested
217 that soil bacteria were one sensitive indicators of soil fertility while Lesueur et al (unpublished data)
218 suggested that indicator based on the fungal communities might not be suitable as microbial indicator
219 for soil quality. On the other hand, appearance of less numerous but strong bands in the fungal DGGE
220 profiles would substantiate the hypothesis that a limited number of dominant and ecologically well-
221 adopted fungal types were present in the soil of the long-term experiment.

222
223 Shifts in bacterial community structure following adaptation of soil management practices have
224 been reported by different studies (Diacono and Montemurro, 2010). The distinct clustering of
225 treatments with and without organic and inorganic inputs suggested a direct effect of organic and
226 inorganic applications on total diversity of bacterial and fungal communities as the percentage
227 similarity in cluster analysis was greater than 18% in the microbial communities. It is always difficult
228 to distinguish between the direct and the indirect effects of an amendment on the behavior of soil

229 microorganisms. However Murphy et al (2007) showed that organic material such as compost or
230 manure are slowly decomposed in the soil, and the continuous release of nutrients can sustain the
231 microbial biomass population for longer periods of time, compared with mineral fertilizers. Our results
232 are in accordance with such study. As a general rule the quantity and quality of organic material added
233 to soils are the major factors in controlling the abundance of different microbial groups and the
234 activity of microorganisms involved in nutrient cycling (Diacono and Montemurro, 2010). Acosta-
235 Martinez et al (2008) demonstrated 30-50% C reduction in soil resulted in a community structure with
236 lower fungal populations and lower enzyme activities compare with undisturbed pastures. Soil Carbon
237 may be one of the key factors influencing soil microbial diversity.

238
239 Considering that agricultural use of inorganic fertilizers unavoidably decrease microbial activity, Ge et
240 al (2010) recommend that combined use of organic manure with inorganic fertilizers should be
241 considered based on the balance between crop demand and soil supply of available nutrients. This is
242 totally in accordance with Manna et al (2007) who showed that if a balance fertilizer, either alone or in
243 combination with manure application, had a positive effect on crop yields, in a cereal based cropping
244 system, application of balanced fertilizer with manure had a great impact on soil fertility improvement.

245

246

247 **Conclusion**

248 Considering the long-term effects of continuous application of farmyard manure, residues and N and P
249 fertilizers we showed that the total diversity of bacterial and fungal communities was considerably
250 affected by the input type. Combination of organic and inorganic resources increased soil bacterial and
251 fungal diversity and resulted in a more even distribution than that in soil treated with inorganic

252 fertilizers and in untreated control soils, these alterations were linked with the availability of organic
253 sources in inorganically treated soils.

254

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261

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