

Ecogeography and utility to plant breeding of the crop wild relatives of sunflower (*Helianthus annuus* L.)

Michael B. Kantar^{1,2*}, Chrystian C. Sosa³, Colin K. Khoury^{3,4}, Nora P. Castañeda-Álvarez³, Harold A. Achicanoy³, Vivian Bernau⁵, Nolan Kane⁶, Laura Marek⁷, Gerald Seiler⁸, Loren Rieseberg^{2,9}

¹Agronomy and Plant Genetics, University of Minnesota, USA, ²Biodiversity Research Centre and Department of Botany, University of British Columbia, Canada, ³DAPA, International Center for Tropical Agriculture (CIAT), Colombia, ⁴Centre for Crop Systems Analysis, Wageningen University, Netherlands, ⁵Department of Horticulture and Crop Science, The Ohio State University, USA, ⁶Department of Ecology and Evolutionary Biology, University of Colorado at Boulder, USA, ⁷North Central Regional Plant Introduction Station, Agronomy Department, Iowa State University and USDA-ARS, USA, ⁸Northern Crop Science Laboratory, USDA-ARS, USA, ⁹Department of Biology, Indiana University, USA

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Provisional

1 **Ecogeography and utility to plant breeding of the crop wild relatives of sunflower**
2 **(*Helianthus annuus* L.)**

3 Michael B. Kantar^{1,2,11}, Chrystian C. Sosa³, Colin K. Khoury^{3,4}, Nora P. Castañeda-Álvarez^{3,5},
4 Harold A. Achicanoy³, Vivian Bernau^{3,6}, Nolan C. Kane⁷, Laura Marek⁸, Gerald Seiler⁹, Loren
5 H. Rieseberg^{1,10}

6 ¹Biodiversity Research Centre and Department of Botany, University of British Columbia, 3529-
7 6270 University Boulevard, Vancouver, British Columbia V6T 1Z4, Canada

8 ²Department of Agronomy and Plant Genetics, University of Minnesota, 411 Borlaug Hall, 1991
9 Upper Buford Circle, St. Paul, MN 55108

10 ³International Center for Tropical Agriculture (CIAT), Km 17, Recta Cali-Palmira, Apartado
11 Aéreo 6713, Cali, Colombia

12 ⁴Centre for Crop Systems Analysis, Wageningen University, Droevendaalsesteeg 1, 6708 PB
13 Wageningen, Netherlands

14 ⁵School of Biosciences, University of Birmingham, Edgbaston, Birmingham B15 2TT, UK

15 ⁶Department of Horticulture and Crop Science, The Ohio State University, 202 Kottman Hall,
16 2021 Coffey Rd, Columbus, Ohio 43210, USA.

17 ⁷Department of Ecology and Evolutionary Biology, University of Colorado at Boulder, Boulder,
18 CO, USA

19 ⁸ North Central Regional Plant Introduction Station, Agronomy Department, Iowa State
20 University and USDA-ARS, Ames, IA 50014, USA

21 ⁹USDA-ARS, Northern Crop Science Laboratory, Fargo, ND 58102, USA

22 ¹⁰Department of Biology, Indiana University, Bloomington, IN 47405, USA

23 ¹¹To whom correspondence should be addressed

24 **Correspondence:**

25 Dr. Michael B. Kantar

26 Department of Agronomy and Plant Genetics

27 University of Minnesota

28 411 Borlaug Hall

29 1991 Upper Buford Circle

30 St. Paul, MN 55108

31 kant0063@umn.edu

32 **Abstract**

33 Crop wild relatives (CWR) are a rich source of genetic diversity for crop improvement.
34 Combining ecogeographic and phylogenetic techniques can inform both conservation and
35 breeding. Geographic occurrence, bioclimatic, and biophysical data were used to predict species
36 distributions, range overlap and niche occupancy in 36 taxa closely related to sunflower
37 (*Helianthus annuus* L.). Taxa lacking comprehensive *ex situ* conservation were identified. The
38 predicted distributions for 36 *Helianthus* taxa identified substantial range overlap and asymmetry
39 and niche conservatism. Specific taxa (e.g., *Helianthus debilis* Nutt., *Helianthus anomalus* Blake,
40 and *Helianthus divaricatus* L.) were identified as targets for traits of interest, particularly for
41 abiotic stress tolerance and adaptation to extreme soil properties. The combination of techniques
42 demonstrates the potential for publicly available ecogeographic and phylogenetic data to
43 facilitate the identification of possible sources of abiotic stress traits for plant breeding programs.
44 Much of the primary genepool (wild *H. annuus*) occurs in extreme environments indicating that
45 introgression of targeted traits may be relatively straightforward. Sister taxa in *Helianthus* have
46 greater range overlap than more distantly related taxa within the genus. This adds to a growing
47 body of literature suggesting that in plants (unlike some animal groups), geographic isolation
48 may not be necessary for speciation.

49 **Key words:** conservation, climate change, crop wild relatives, ecological niche modeling, plant
50 breeding, plant genetic resources, publicly available data sources

51 **Introduction**

52 Plant genetic resources represent the biological foundation for maintaining and improving
53 crop productivity having played a central role in crop development from antiquity (Porter *et al.*,
54 2014). Crop wild relatives (CWR) are an important source of useful traits for plant breeding
55 (Hoisington *et al.*, 1999; Hajjar & Hodgkin, 2007). With the world's population projected to
56 increase the need to produce more food while using fewer natural resource inputs under
57 increasingly stochastic climatic conditions is a major challenge (Challinor *et al.*, 2014; Butler &
58 Huybers, 2013). CWR conservation and utilization focusing on the use of improving
59 technologies (high throughput phenotyping, genotyping, and geographical information systems),
60 has been proposed as a way to acquire a greater knowledge of conservation needs and lead to
61 more targeted use of CWR germplasm (Khoury *et al.*, 2010; Cabrera-Bosquet *et al.*, 2012;
62 McCouch *et al.*, 2013). Targeted collecting for *ex situ* conservation has become a priority as
63 rapid changes in both climate and land use patterns increasingly threaten CWR in their natural
64 habitats (Jarvis *et al.*, 2008; McCouch *et al.*, 2013).

65 Crop wild relatives have traditionally been categorized based on crossing relationships
66 with domesticates; the primary germplasm contains no crossing barriers, the secondary contains
67 some meiotic abnormalities, and the tertiary requires special techniques such as embryo rescue
68 (Harlan & De Wet, 1971; Harlan, 1976). Such classifications may be supplemented by
69 molecular, bioclimatic and biophysical data to aid in the identification of candidate taxa for
70 breeding, although such efforts have been constrained by challenges in comprehensively
71 generating and integrating these data (Ricklefs & Jenkins, 2011).

72 The genus *Helianthus* L. contains 52 species comprising 67 taxa (Schilling, 2006;
73 Stebbins *et al.*, 2013). Native to North America, the taxa occupy a variety of habitats ranging

74 from open plains to salt marshes (Kane *et al.*, 2013; Seiler & Marek, 2011). Sunflower
75 (*Helianthus annuus* L.) is the most economically important species from the genus, with ~26
76 million hectares in production worldwide and a substantial private sector breeding effort,
77 particularly for oil production (FAOSTAT, 2013). Domesticated approximately 4000 years ago
78 in east central North America, sunflower has a typical domestication syndrome; i.e., it does not
79 branch, does not have seed dormancy, has a predictable flowering time, and does not shatter
80 (Harlan *et al.*, 1973; Harter *et al.*, 2004; Blackman *et al.* 2011). The crop has undergone both
81 selection and genetic drift during domestication and improvement, which has reduced genetic
82 diversity (Liu & Burke 2006; Tang & Knapp 2003), with modern cultivars retaining 50-67% of
83 the diversity present in wild *H. annuus* populations (Kolkman *et al.* 2007; Mandel *et al.*, 2011).

84 Sunflower has often utilized CWR in breeding efforts, with many of the taxa hybridizing
85 well with the crop (Table S1; Table 1) (Long *et al.*, 1960; Chandler *et al.*, 1986). Despite the
86 historical use, CWR of sunflower are considered to be relatively untapped, particularly in regard
87 to adaptation to abiotic stresses. To contribute to an enhanced understanding of the CWR of
88 sunflower, this studies' objectives were to 1) create geographical distribution models for 36
89 CWR taxa, and 2) explore niche habitation through comparisons of ecogeographic and
90 phylogenetic data, to identify taxa occurring in extreme environments of potential interest to
91 sunflower breeding.

92 **Materials and Methods**

93 ***Species distribution modeling***

94 A modified gap analysis (Ramírez-Villegas *et al.*, 2010) was used to determine the conservation
95 status of 36 taxa within *Helianthus* selected based upon their potential to provide useful traits for
96 sunflower breeding. Briefly, 1) target taxa were identified, and geographic occurrence data were

97 gathered and verified, 2) the overall representation of CWR in germplasm collections was
98 estimated, 3) potential distribution models were produced for taxa with sufficient samples with
99 coordinates, 4) the geographic and ecological representation of germplasm collections were
100 assessed for each taxon by comparing potential distribution models to existing germplasm
101 collection locations, 5) taxa were prioritized for further collecting based upon the average of
102 their overall, geographic, and ecological coverage results, and 6) gap analysis results were
103 correlated with the subjective assessments of collection priorities from crop experts.

104 The selection of taxa for analysis was based on membership within the primary or
105 secondary gene pools of sunflower (Vincent *et al.*, 2013) with the addition of all taxa from the
106 tertiary gene pool indicated in publications to be confirmed or potential trait donors (Table S1). A
107 total of 12,737 occurrence records for the 36 taxa, sourced from 31 herbaria and five genebanks,
108 were used for distribution models and conservation analysis (Table S2), including 4,705 records
109 with geographic coordinates. The overall representation of taxa in genebank collections was
110 estimated using the ‘Sampling Representativeness Score’ (SRS), calculated as the number of
111 germplasm samples (GS) divided by the total number of samples (GS plus reference records).
112 After eliminating duplicate records, potential distributions were calculated using Maxent
113 (Phillips *et al.*, 2006), with a k-5 cross-validation option and 10,000 background points for model
114 training over North America (Phillips, 2008; VanDerWal *et al.*, 2009). We included nineteen
115 bioclimatic variables derived from the WorldClim database (Nix, 1986; Hijmans *et al.*, 2005a;
116 Hijmans *et al.*, 2005b), seven biophysical variables from the ISRIC – World Soil Information
117 database (<http://soilgrids1km.isric.org>) at a resolution of 2.5 arc-minutes, and the occurrence
118 information (coordinates) for each taxon as inputs (Table S3). For edaphic data we calculated a
119 weighted mean from five depths (0 to 5 cm, 5 to 15 cm, 15 to 30 cm, 30 to 60 cm, 60 to 100 cm) to

120 generate a single value for the first meter of soil for each layer, and then resampled the data from
121 1 arc minutes to 2.5 arc minutes resolution to match the WorldClim dataset, using the raster
122 package in R and ArcGIS Desktop 10.1 (Hengl *et al.*, 2014). Distributions were further restricted
123 by applying a taxon independent threshold, based on the Receiver Operating Characteristic
124 (ROC) curve (Liu *et al.*, 2005). GRIN distribution data was used to ensure that taxa distributions
125 were not overinflated beyond known native boundaries (GRIN, 2012). Soil cover data from
126 GlobCover 2009 (Global Land Cover Map) (<http://due.esrin.esa.int/globcover/>) further refined
127 the maxent outputs and collecting maps by excluding urban areas, water bodies, bare areas, and
128 permanent snow and ice regions.

129 Potential distribution models were considered accurate if they complied with the
130 following conditions: i) 5-fold average area under the test ROC curve (ATAUC) is greater than
131 0.7, ii) the standard deviation of ATAUC (STAUC) is less than 0.15, and iii) At least 10% of
132 grids for each model has standard deviation less than 0.15 (ASD15). For taxa whose Maxent
133 model did not comply, potential distributions were estimated by forming a circular buffer of 50
134 km around each occurrence point for each species.

135 Geographic representativeness of taxa in genebank collections was calculated using the
136 ‘Geographic Representativeness Score’ (GRS), comparing the spatial overlap of a circular buffer
137 surrounding each accession record (50 Km radius as described in Hijmans *et al.*, 2001) against
138 the potential distribution of the taxon. Ecological gaps in genebank collections were calculated
139 using the ‘Ecological Representativeness Score’ (ERS), calculated by comparing records to the
140 full environmental range of the modeled taxon across ecosystem types (Olson *et al.*, 2001). The
141 overall priority for further collecting for *ex situ* conservation for each taxon was determined by
142 averaging the SRS, GRS, and ERS with equal weight to obtain a final prioritization score (FPS),

143 classified according to the following ranges: 1., high priority (FPS between 0 and 3); 2., medium
144 priority (FPS between 3.01 and 5); 3., low priority (FPS between 5.01 and 7.5); and 4., and well
145 conserved taxa (FPS between 7.51 and 10).

146 ***Expert evaluation of conservation assessment results***

147 Predicted taxon distributions based on genebank and herbarium records were compared to the
148 knowledge of four crop experts with experience with *Helianthus* distributions, systematics,
149 conservation and diversity. *Helianthus* experts were asked to evaluate of the adequacy of
150 germplasm collections per species based on their knowledge of total accessions conserved,
151 geographic and environmental gaps. This assessment was given an expert priority score (EPS),
152 analogous to the FPS score. A second score was generated, the contextual EPS, which based on
153 additional knowledge such as *in situ* threats and utility to crop breeding. After initial evaluation
154 the experts were asked to review the quantitative results, occurrence data, potential distribution
155 models, and maps of collecting priorities. Following expert input, occurrence data were refined
156 through elimination of incorrect points and adjustment native areas. Potential distribution
157 modeling and gap analyses were then conducted using refined datasets to create more accurate
158 species distribution maps. Potential zones for collecting were identified for each high priority
159 taxon, and then combined to create maps depicting areas where multiple taxa of high priority for
160 conservation could be collected.

161 ***Ecogeographic niche overlap and phylogenetic analyses***

162 Potential distribution probability outputs were used when Maxent models performed well
163 and CA50 sample buffers when Maxent models did not pass the validation criteria, to calculate
164 niche overlap based on Schoener's D and Hellinger's I as outlined in Warren *et al.* 2008, and
165 implemented in the R package Phyloclim (Heibl, 2011). Both indices utilize probability

166 distributions in geographic space, with statistics ranging from 0 (no niche overlap) to 1
167 (complete niche overlap). First pairwise niche overlap was examined, then niche overlap
168 between allopatric/sympatric taxa separately, annual/perennial taxa separately, and lastly
169 allopatric/sympatric sister taxa. Geographic range overlap for all pairwise combinations (630
170 comparisons) was calculated in two ways, with respect to the larger range $[(2 * \text{number of shared}$
171 $\text{grid cells}) / (\text{number of grid cells in taxa A} + \text{number of grid cells in taxa B})]$ and with respect to
172 the smaller range $[(2 * \text{number of shared grid cells}) / (\text{Total number of grid cells in taxa A} + \text{Total}$
173 $\text{number of grid cells in taxa B})] / (\text{Total potential number of shared grid cells}) [2 * \text{total number of}$
174 $\text{grid cells in species with the smaller range}] / (\text{Total number of species A} + \text{Total number of}$
175 $\text{species B})]$.

176 Principal component analyses (PCA) were used to assess the importance of
177 ecogeographic variables (Table S3) to variation in occurrence data of distribution models per
178 taxon. A hierarchical cluster of principal components (HCPC) identified climatic clusters using
179 R package FactoMineR (Husson *et al.*, 2014). Boxplots for each bioclimatic and biophysical
180 layer were created based on occurrence data points (Fig. S1). Ecogeographic variables for
181 cultivated sunflower were extracted from the area of species distribution maps (Monfreda *et al.*,
182 2008) at a resolution of 5 arc-minutes, with a random sample of 1,000 points weighted by
183 harvested area taken from major production regions.

184 We downloaded the publically available 18S-26S Ribosomal DNA sequence from the
185 external transcribed spacer (ETS) from GenBank (NCBI-<http://www.ncbi.nlm.nih.gov/>) for 28 of
186 the 36 *Helianthus* taxa, aligned the sequences using ClustalW, and constructed a maximum
187 likelihood phylogeny with 1000 bootstrap replications, using MEGA6 with a Jukes-Cantor
188 nucleotide substitution model (Tamura *et al.*, 2013). We performed a Mantel test in R utilizing

189 the ade4 package to explore the relationship between geography and genetics (Dray & Dufour,
190 2007). We estimated phylogenetic signal of individual ecogeographic traits utilizing Blomberg's
191 K (Blomberg, et al, 2003), using the multiphylosignal command with 1000 permutations in
192 Picante (Kembel *et al.*, 2010).

193 **Results**

194 ***Geographic distributions of sunflower crop wild relatives***

195 Predicted distribution maps were produced for 36 *Helianthus* taxa, along with taxon
196 richness and collecting hotspot maps (Fig. 2; Fig. S2). Thirty of the 36 taxa (83%) produced
197 valid maxent models with utilization of soil pH and percent sand greatly improving the accuracy
198 of distribution models, as assessed by expert opinion (Fig. 3). Five hotspots (areas of high taxon-
199 level diversity) were identified in the USA, including the southeastern gulf coast, the south-
200 central, the midwest, the north central, and the central east coast (Fig. 2a). Our results suggest
201 that half of the 36 taxa are in urgent need of further collecting (high priority species – HPS),
202 along with 28% in moderate need (medium priority species – MPS), 6% of low priority (LPS),
203 and 17% that are well represented in existing germplasm collections and thus do not require
204 urgent additional collecting (Table 1). While the primary gene pool taxa has been well collected,
205 only 10% of the taxa in the secondary gene pool are well represented across their geographic,
206 climatic, and edaphic ranges. Likewise, only 7% of taxa in the tertiary gene pool were assessed as
207 well-conserved (Fig. 1; Table 1). These results contrasted with those of expert reviewers, who
208 classified more species as LPS. The discrepancy between the results and expert opinion was due
209 in part to overly optimistic distribution models regarding likelihood of occurrence, in comparison
210 to expert realities of existence of populations in these regions. Additionally, experts assessed
211 some taxa, such as *Helianthus debilis* ssp. *cucumerifolius*, at lower priority because distributions

212 have expanded recently as weedy populations invade new areas, and such regions were not
213 considered by the experts as of particular priority.

214 *Ecological niches of sunflower crop wild relatives*

215 Three ecogeographic clusters differentiate the taxa, with the first three PCs accounted for
216 74.3% of the variation (Fig. 3b; Table S4). Clusters broadly corresponded to plain, desert, and
217 woodland ecosystems (Table 1). Cluster one was mostly composed of the secondary germplasm
218 and differentiated by temperature, while cluster two was mostly the tertiary germplasm and
219 differentiated by precipitation. Cluster three was differentiated by soil and was evenly split
220 between the secondary and tertiary germplasm (Table S3). It is important to note that PCA can
221 increase type one error, so ecological niches must be carefully examined and validated (Revel,
222 2009; Uyeda et al., 2015). Schoener's D and Hellinger's I identified substantial niche overlap
223 with few taxa showing niche divergence (Fig. 3; Table 1).

224 Potential geographic distributions of crop wild relative taxa were examined for overlap
225 with wild *H. annuus* (Fig. S1); most (81%) taxa exhibited some geographic range overlap with
226 *H. annuus* (Table 1). Among CWR taxa, 39% of pairwise comparisons had overlapping
227 geographic distributions (sympatry), while 61% were allopatric (Table S5; Fig. S3). Eight of the
228 twelve sister taxa pairs among the CWR showed some level of sympatry (Table S6). There was
229 considerable range asymmetry between taxa (Fig. S1), with the amount of overlap depending on
230 the direction of the comparison, where the smaller range showed 26% more overlap on average
231 than the larger range (Table S5).

232 There was general niche conservatism even for sister-taxa (Fig. 3; Table 2). While
233 ecogeographic niches were fairly similar for many variables, occasionally there was substantial
234 divergence (Fig. 4; Fig. S1). Phylogenetic niche conservatism was found in ~54% of variables

235 (Fig. 5). Divergence was found in several soil variables suggesting an important role of soil in
236 *Helianthus* diversification. A Mantel's test using Mahalanobis distance ($r=0.1423$, $p=0.01$),
237 indicated that taxa that are geographically close are generally more closely related genetically.
238 Notable exceptions to this were *H. maximilliani*, *H. grosseserratus*, and *H. giganteus*, which are
239 sympatric with *H. annuus*, but are distantly related.

240 **Discussion**

241 There has been increased effort to digitize data related to plant species in general and
242 CWR in particular. The public databases (GBIF, ISRIC, WorldClim, National Germplasm
243 repositories, DivSeek) that archive these data are an increasingly important tool to
244 conservationists, evolutionary biologists and plant breeders. Utilizing public data can reduce the
245 research costs in terms of people hours and consumables to achieve desired environmental and
246 food production goals. Exploring public databases can provide a targeted way to identify
247 accessions for introgression that can then be used to validate predicted extreme variation. This
248 may be a way to more quickly utilize germplasm collections and provide a link to international
249 initiatives aimed at facilitating more use of plant genetic resources (www.DivSeek.org). Here we
250 have used geographic occurrence, bioclimatic, and biophysical data to predict species
251 distributions, range overlap and niche occupancy in 36 *Helianthus* taxa that are cross-compatible
252 with cultivated sunflower and thus likely to be useful in crop breeding. As discussed briefly
253 below, our results not only have implications for conservation genetics and breeding in
254 *Helianthus*, but they also impact our understanding of the role of geography in the origin of
255 species in this group.

256 ***Implications for conservation and plant breeding***

257 Our approach is both new and complementary to previous work on *Helianthus* species
258 distributions and CWR in the literature (Thompson *et al.*, 1981; Rogers *et al.*, 1982). The method

259 of constraining ranges to known native distributions may have limited our identification of some
260 the extreme variation. Despite this, many taxa that diverge ecologically from cultivated
261 sunflower were identified (Fig. 4; Table 1). It was also possible to identify extreme populations
262 within taxa that showed potential adaptation to different ecological niches.

263 Taxa with larger ranges tend to have greater resilience to changes in environmental
264 conditions than taxa with more limited distributions (Sheth & Angert, 2014; Sexton *et al.*, 2014).
265 Thus, the latter may be considered a primary priority for conservation. Several taxa have
266 expanded far beyond their historical ranges, including *H. annuus*, *H. petiolaris* Nutt., *H.*
267 *argophyllus* Torrey & Gray, *H. giganteus* L. and *H. tuberosus* L.. While taxa from the non-native
268 parts of their ranges have not been prioritized, existing accessions from such ranges are
269 acknowledged, and may be worthwhile for exploration for traits useful in crop breeding.

270 Clustering of CWR by environmental variables has great utility by allowing genetic
271 resources to be exploited in a more targeted manner. For example, with respect to soil pH the
272 taxa *H. atrorubens*, *H. resinosus*, and *H. deserticola* occupy different ecological space from
273 cultivated *H. annuus* (Fig. 4). These taxa represent potential candidates for tolerance to acid or
274 alkaline soils, particularly to improve the ability of the crop to accumulate heavy metals for
275 phytoremediation (Fassler *et al.*, 2010). Surprisingly, when examining the properties of the
276 primary, secondary and tertiary germplasm, often extreme profiles are found in the primary
277 germplasm. This is fortuitous since introgression from primary germplasm is more likely to be
278 successful (Fig. 4; Fig. S1; Table S7). Approximately 650 wild *H. annuus* accessions are
279 conserved in genebanks which occur outside the ecological parameters of the cultivar (Table S7).
280 The general reduction of environmental diversity occupied by the cultivated sunflower relative to
281 wild *H. annuus* may indicate the reduction in genetic diversity occurring through domestication.

282 Recent advances in plant and animal breeding (e.g. marker assisted selection, genomic
283 selection) have been facilitated by low cost molecular marker technologies resulting in new tools
284 that can be used to broaden the genetic base in crops (Tester & Langridge, 2010). These methods
285 can shorten breeding cycles, increasing genetic gain per unit time, and allow for wider crosses to
286 be utilized by minimizing linkage drag (Bernardo, 2008). The recent development of genome
287 wide marker sets (Bowers *et al.*, 2012; Renaut *et al.* 2013) and release of the *H. annuus* genome
288 (Kane *et al.*, 2011; <http://www.sunflowergenome.org>) facilitate the use of marker assisted
289 selection (Iftekharuddaula *et al.*, 2011) by decreasing costs and increasing data resolution.
290 Further, if germplasm collections are genotyped, these data can be used to associate particular
291 allelic variants with environmental adaptation (Fang *et al.*, 2014).

292 ***Range overlap of wild relatives of sunflower***

293 Sister species in *Helianthus* often have overlapping ranges, an observation that is
294 consistent with sympatric and “budding” speciation (parapatric or peripheral range speciation).
295 Substantial range asymmetry among some (but not all) sister species is also consistent with a
296 budding speciation scenario (Table S6). The amount of range overlap between sister taxa in
297 *Helianthus* is similar to recent reports from other plant genera, but different from many animal
298 groups, where allopatry tends to be the rule in speciation (Mayr, 1954; Soltis *et al.*, 2004;
299 Quenouille *et al.*, 2011; Anacker & Strauss, 2014). This may suggest that geographic isolation is
300 less critical to plant than animal speciation, perhaps because of the low vagility of many plant
301 species.

302 Unlike sympatric congeners in other plant groups (Grossenbacher *et al.*, 2014; Anacker &
303 Strauss, 2014), *Helianthus* sister taxa typically lack strong ecological divergence. This
304 observation is inconsistent with most models of speciation involving gene flow, which assume

305 divergent ecological selection (Via, 2009). Possibly, our analyses lacked sufficient resolution or
306 focus on key ecological attributes to detect real differences between the ecological niches of
307 these species. For example, it is possible that there has been pollinator and phenological
308 divergence between sister species that was not included in our analyses. Alternatively, local
309 niche differences between sympatric populations may have been masked by substantial
310 ecological heterogeneity among populations of the more widely ranging species. Additionally,
311 the approach used was designed to analyze potential habitat in the historical, native range, rather
312 than recent range expansions, which in many cases may be recent introductions facilitated by
313 humans, perhaps accounting for observations of limited ecological divergence.

314 Our analyses imply that many *Helianthus* taxa have similar ecological niches and exhibit
315 niche conservatism. Under niche conservatism, greater allopatric and parapatric speciation is
316 predicted, as habitat fragmentation is expected to contribute to reproductive isolation (Loera *et*
317 *al.*, 2012). While such a speciation strategy would be surprising given the overlap in geographic
318 range of sister species within *Helianthus*, this trend has been observed in North American
319 *Ephedra* (Loera *et al.*, 2012). That larger amount of niche conservatism observed here than in
320 other systems may be due to properties of the K-statistic, which can have inflated values in
321 polyphyletic phylogenies and in the presence of incomplete lineage sorting, both of which occur
322 in *Helianthus* (Rosenthal *et al.*, 2002; Gross & Rieseberg, 2005; Horandl & Stuessy, 2010;
323 Davies *et al.*, 2012).

324 **Conclusions**

325 Using a combination of gap analysis, environmental niche modeling and phylogenetic
326 approaches 36 CWR of sunflower were examined. Taxa that are under-represented in germplasm
327 collections as well as species and populations inhabiting environmental niches with extreme

328 phenotypes that may possess traits of value to crop improvement were identified. In *Helianthus*,
329 sister taxa appear to occur more frequently in sympatry than allopatry, possibly suggesting that
330 speciation may occur in the presence of gene flow. Finally, much of the primary gene pool occurs
331 in extreme environments indicating that utilization of wild *H. annuus* for the breeding of abiotic
332 stress tolerance may produce quick gains with minimal effort.

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Table 1. Taxa examined in this study, recommendation, position in germplasm, environmental cluster, life history, and potential extreme characteristics.

<u>Taxa</u>	<u>Recommendation for Collection</u>	<u>Position in Germplasm</u>	<u>Range overlap with <i>H. annuus</i></u>	<u>Environmental Cluster Assignment</u>	<u>Life History</u>	<u>Potential Extreme Characteristics Based on Different Ecological Niche Relative to <i>H. annuus</i></u>
<i>H. annuus</i> (wild)	Assessed to be well represented	Primary	NA	Cluster 1	Annual	NA
<i>H. anomalus</i>	High priority	Secondary	Utah New Mexico	Cluster 3	Annual	Low precipitation tolerance Tolerance to high pH
<i>H. argophyllus</i>	Medium priority	Secondary	Texas	Cluster 1	Annual	High temperature tolerance Tolerance to high clay content
<i>H. arizonensis</i>	Medium priority	Tertiary	Arizona New Mexico	Cluster 3	Perennial	Response to stochastic climate Low precipitation tolerance Tolerance to low bulk density
<i>H. atrorubens</i>	Medium priority	Tertiary	No overlap	Cluster 2	Perennial	Tolerance to low Cation-exchange capacity Tolerance of high precipitation Tolerance to low pH
<i>H. bolanderi</i>	High priority	Secondary	California	Cluster 1	Annual	Tolerance to erratic precipitation Low precipitation tolerance
<i>H. debilis</i> subsp. <i>cucmerifolius</i>	High priority	Secondary	East Texas	Cluster 2	Annual	High temperature tolerance
<i>H. debilis</i> subsp. <i>debilis</i>	Medium priority	Secondary	No overlap	Cluster 2	Annual	High temperature tolerance Tolerance of high precipitation Tolerance to low clay content
<i>H. debilis</i> subsp. <i>silvestris</i>	Medium priority	Secondary	No overlap	Cluster 2	Annual	Tolerance to high clay content
<i>H. debilis</i> subsp. <i>tardiflorus</i>	Assessed to be well represented	Secondary	No overlap	Cluster 2	Annual	Tolerance of high precipitation Tolerance to low clay content
<i>H. debilis</i> subsp. <i>vestitus</i>	Low priority	Secondary	No overlap	Cluster 2	Annual	High temperature tolerance Tolerance of high precipitation Tolerance to low clay content
<i>H. deserticola</i>	High priority	Secondary	Nevada Utah New Mexico	Cluster 3	Annual	Response to stochastic climate Low precipitation tolerance
<i>H. divaricatus</i>	High priority	Tertiary	Central US	Cluster 2	Perennial	Perennial habit Tolerance to low pH
<i>H. exilis</i>	Medium priority	Secondary	California	Cluster 1	Annual	Tolerance to erratic precipitation Low precipitation tolerance Low bulk density
<i>H. giganteus</i>	High priority	Tertiary	No overlap	Cluster 2	Perennial	Tolerance of high precipitation
<i>H. grosseserratus</i>	Medium priority	Tertiary	Central US	Cluster 3	Perennial	Tolerance to erratic temperature
<i>H. hirsutus</i>	High priority	Tertiary	Central US	Cluster 2	Perennial	Tolerance to low pH
<i>H. maximilliani</i>	High priority	Tertiary	Central US	Cluster 3	Perennial	Low temperature tolerance Tolerance to erratic temperature
<i>H. neglectus</i>	Assessed to be well represented	Secondary	New Mexico	Cluster 1	Annual	Low organic carbon content
<i>H. niveus</i> subsp. <i>canescens</i>	High priority	Secondary	California Arizona New Mexico	Cluster 1	Annual Rarely Perennial	High temperature tolerance Low precipitation tolerance
<i>H. niveus</i> subsp. <i>niveus</i>	High priority	Secondary	Baja California	Cluster 1	Perennial	Low precipitation tolerance
<i>H. niveus</i> subsp. <i>tephrodes</i>	High priority	Secondary		Cluster 1	Perennial	High temperature tolerance low Precipitation tolerance

			California, Mexico (Sonora)		Sometime Annual	
<i>H. paradoxus</i>	Assessed to be well represented	Secondary	Texas, New Mexico	Cluster 1	Annual	Low organic carbon content
<i>H. pauciflorus</i> subsp. <i>pauciflorus</i>	High priority	Tertiary	Central US	Cluster 3	Perennial	Tolerance to erratic temperature
<i>H. pauciflorus</i> subsp. <i>subrhomboideus</i>	High priority	Tertiary	Central US	Cluster 3	Perennial	Low temperature tolerance Tolerance to erratic temperature
<i>H. petiolaris</i> subsp. <i>fallax</i>	High priority	Secondary	Western US	Cluster 3	Annual	Tolerance to erratic temperature
<i>H. petiolaris</i> subsp. <i>petiolaris</i>	High priority	Secondary	Central US	Cluster 3	Annual	Tolerance to erratic temperature Low temperature tolerance
<i>H. praecox</i> subsp. <i>hirtus</i>	Assessed to be well represented	Secondary	West Texas	Cluster 1	Annual	High temperature tolerance
<i>H. praecox</i> subsp. <i>praecox</i>	Assessed to be well represented	Secondary	East Texas	Cluster 2	Annual	Tolerance to erratic temperature
<i>H. praecox</i> subsp. <i>runyonii</i>	Low priority	Secondary	Texas	Cluster 1	Annual	Tolerance of high bulk density
<i>H. resinus</i>	Medium priority	Tertiary	No overlap	Cluster 2	Perennial	Tolerance of high precipitation Tolerance to low Cation exchange capacity Tolerance to low pH
<i>H. salicifolius</i>	Medium priority	Tertiary	Oklahoma Kansas Arkansas Missouri	Cluster 3	Perennial	Tolerance to high clay content
<i>H. silphioides</i>	Assessed to be well represented	Tertiary	Oklahoma Arkansas Missouri	Cluster 2	Perennial	Tolerance to low cation-exchange capacity Tolerance to low pH
<i>H. strumosus</i>	High priority	Tertiary	Central US	Cluster 2	Perennial	Tolerance of high precipitation
<i>H. tuberosus</i>	Medium priority	Secondary	Central US	Cluster 2	Perennial	Low temperature tolerance
<i>H. winteri</i>	High priority	Primary	California	Cluster 1	Perennial	High temperature tolerance

534 **Table 2. Environmental Niche occupancy based on Schoener's D (1968) and a modified**
 535 **Hellinger's I (Warren *et al.*, 2008).**

	<u>Perfect Overlap</u>	<u>D or I Greater than 0.5</u>	<u>D or I Less than 0.2</u>
	<u>(%)</u>	<u>(%)</u>	<u>(%, Divergent Niche)</u>
All taxa	36.9	69.4	4.7
Annual taxa	32.2	36.6	6.6
Perennial taxa	19.8	85.7	2.2
Allopatric taxa	54.2	62.5	4.3
Sympatric taxa	3.3	83.3	2.6
Sister taxa	33.3	57.7	2.6

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Provisional

539 **Figure Legends**

540 Fig. 1 Synthesis of gap analysis results and expert assessments for each of the 36 *Helianthus*
541 CWR taxa surveyed. Taxa are listed by descending priority for further collecting by category:
542 HPS, high priority taxa; MPS, medium priority taxa; LPS, low priority taxa; NFCR, no further
543 collecting recommended. The final priority scores (FPS, black circle) is the mean of the sampling
544 representativeness score (SRS, blue circle), geographic representativeness score (GRS, red
545 circle), and ecological representativeness score (ERS, green circle).

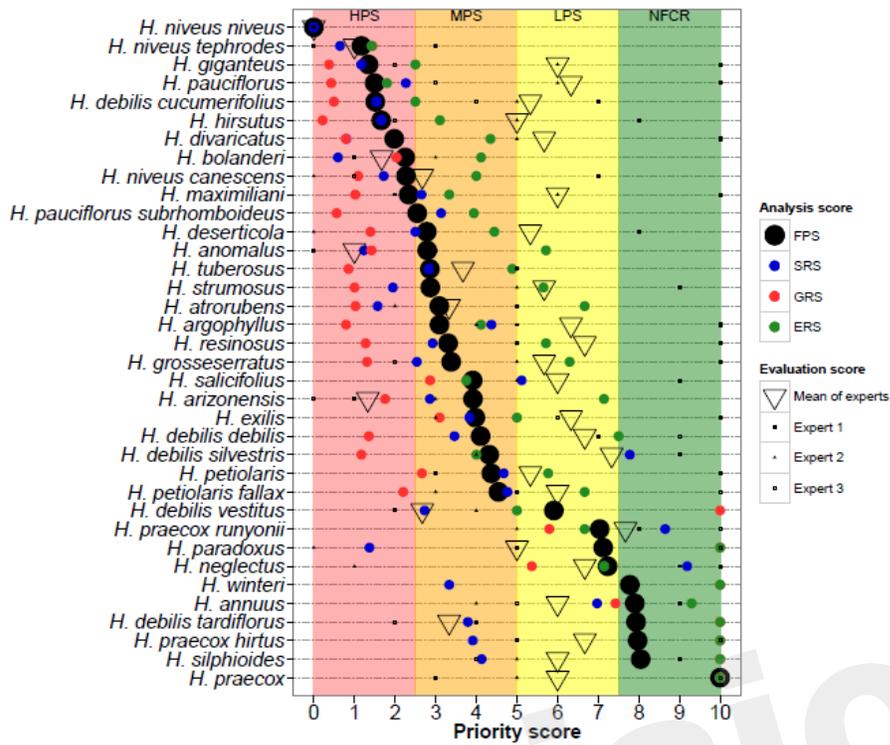
546 Fig. 2 Map of North America showing A) taxon richness of sunflower and B) hotspots for further
547 collecting of high priority taxa.

548 Fig. 3 Geographic niche overlap based on bioclimatic and biophysical variables, both calculated
549 by D (above diagonal) and I (below diagonal) . Taxa are grouped by the phylogenetic
550 relationship identified from the ETS sequences retrieved from NCBI. Values closer to 0 (no
551 overlap = niche divergence) are purple while values closer to 1 (complete overlap = niche
552 convergence) are orange; B) Occurrence points for each taxa grouped based on the first three
553 principle components of biophysical and bioclimatic variables. Clusters share homogeneous
554 bioclimatic and biophysical conditions.

555 Fig. 4 Climatic niches for A) mean diurnal range and annual precipitation, B) Soil pH and mean
556 annual precipitation, C) mean diurnal range and annual precipitation. Niches per taxa represent
557 the middle 90% of occurrence points, i.e., 10% outliers are not included. Red boxes show the
558 niche of wild *H. annuus* and black boxes show the niche of cultivated *H. annuus* in North
559 America.

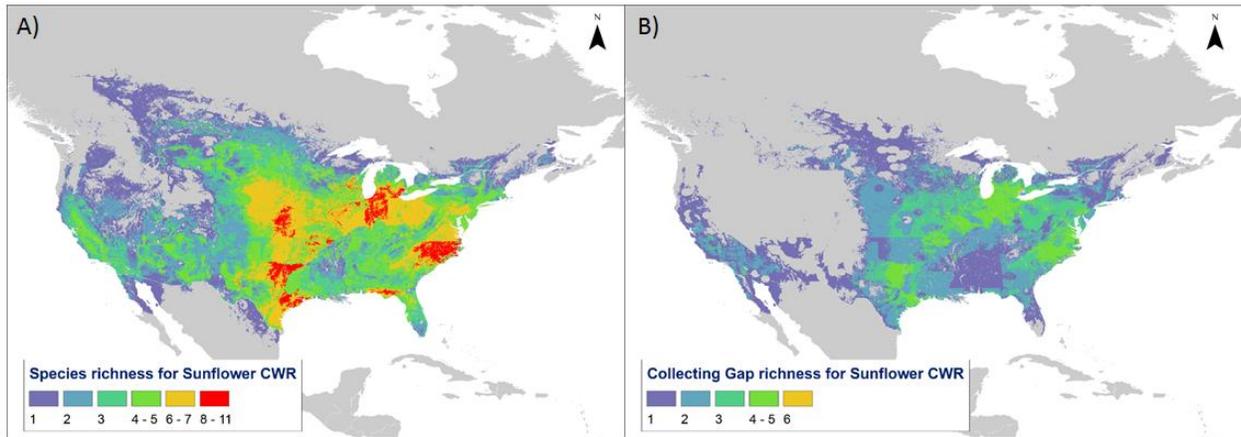
560 Fig. 5 Test of phylogenetic signal utilizing the K for 25 of 36 taxa analyzed with complete
561 genetic and environmental information (Blomberg, et al, 2003). K measures phylogenetic signal
562 in traits, where K values below 1 indicates low dependence of traits on evolutionary history (not
563 conserved between taxa) and K values above 1 indicates trait conservation over evolutionary
564 history (traits conserved over evolutionary time). *indicates K significantly greater than 1 ($p <$
565 0.05).
566

567 **Fig. 1**



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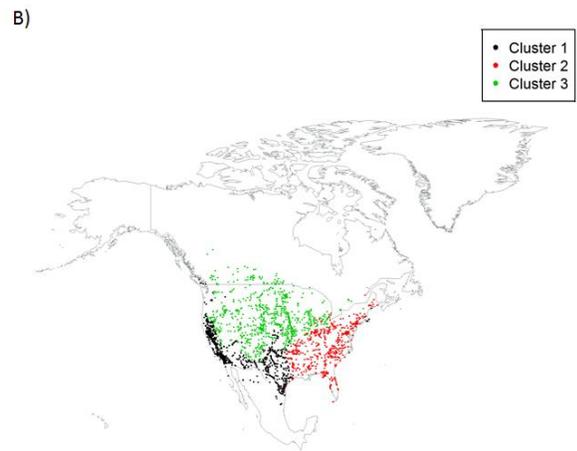
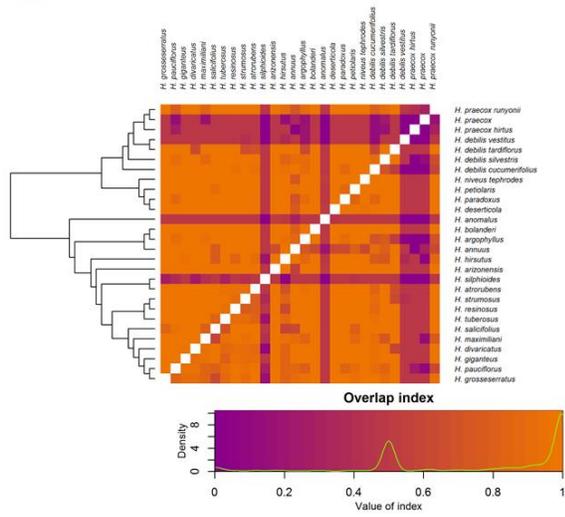
570 **Fig. 2.**
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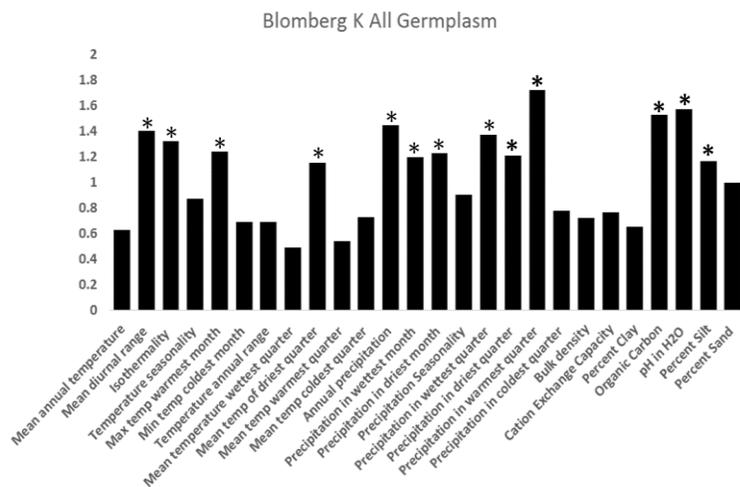
574 **Fig. 3**
A)



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579 **Fig. 5**



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581 **Supplementary information**

582

583 Table S1. *Helianthus* taxa which have provided useful traits for cultivated sunflower.

584 Table S2. Name and location of the 31 herbaria and five germplasm Institutes from which taxa
585 data were sourced.

586 Table S3. Bioclimatic and biophysical variables examined and correlation between climatic
587 variables and selected principal components.

588 Table S4. Bioclimatic and biophysical variables partitioned into clusters using the R package
589 FactoMineR variables. All of the cluster 1 variables are related to temperature and cluster 1 can
590 be defined by dry climatic conditions. Cluster 2 is defined by precipitation variables, and is
591 associated with humid climatic conditions and high soil organic matter. Cluster 3 contains a
592 combination of soil and temperature variables. This cluster has soils with higher than average silt
593 content, a higher capacity for cation exchange, neutral pH, and higher soil porosity.

594 Table S5. A) Geographic overlap as determined with respect to the smaller (minor range) in the
595 bottom left, and larger range (major) in the top right. B) Difference between minor and major
596 range overlap. Red indicates no geographic overlap, white indicates a small amount of overlap
597 and blue indicates a larger amount of overlap.

598 Table S6. Geographic overlap of 12 sister taxa pairs present in our data represented as percent of
599 shared grid cells.

600 Table S7. Populations of wild *H. annuus* that are outliers relative to domestic *H. annuus* so that
601 they may be useful for abiotic stress breeding, (yellow indicates lower than 2.5% of the domestic
602 *H. annuus* distribution, blue indicates outside the 97.5% of the domestic *H. annuus* distribution).

603 Fig. S1 Climatic niches of *Helianthus* species per bioclimatic variable.

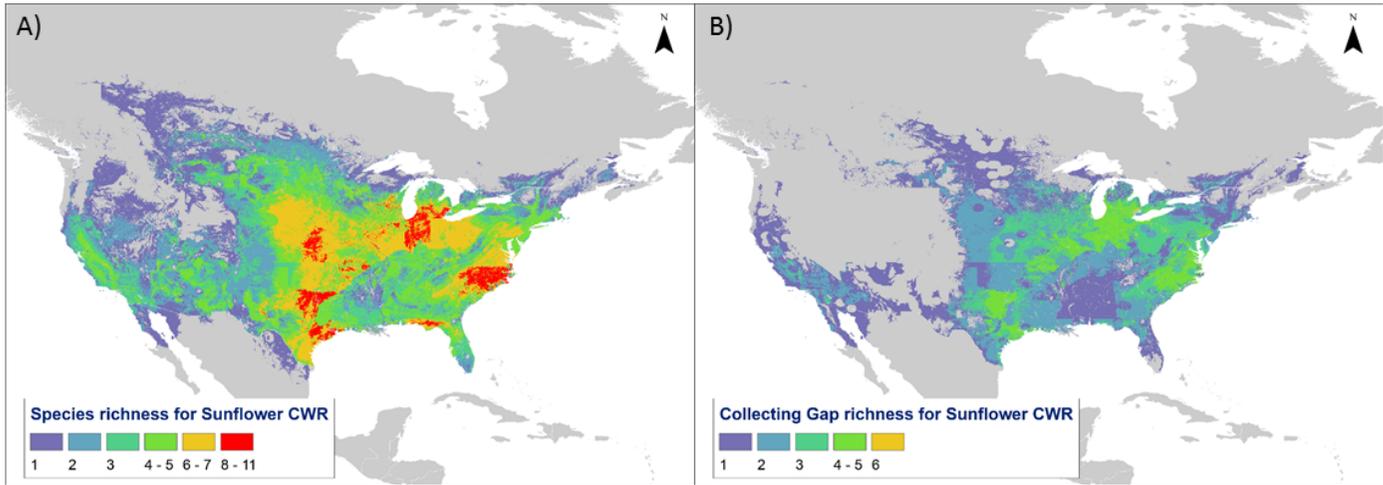
604 Fig. S2 Species distribution maps for the 36 *Helianthus* taxa examined in this study.

605 Fig. S3 Heat map of geographic overlap as determined with respect to the smaller (minor range)
606 in the bottom left, and larger range (major) in the top right. Red indicates no geographic overlap,
607 white indicates a small amount of overlap and blue indicates a larger amount of overlap.

608 Fig. S4 Predicted Niche Occupancy (PNO) for all 19 bioclimatic and 7 biophysical variables.
609 Horizontal axes represent the bioclim parameter space divided into 50 equally spaced bins;
610 vertical axes denote the total suitability of the mean annual temperature index of each species
611 over its entire geographic range. Overlapping peaks of PNO profiles indicate similar tolerances,
612 while the overall breadth of the profile denotes the degree of specificity in tolerance. Black
613 profiles indicate the primary germplasm, red indicates the secondary germplasm pool, blue
614 indicates the tertiary germplasm pool.

Figure 2.TIF

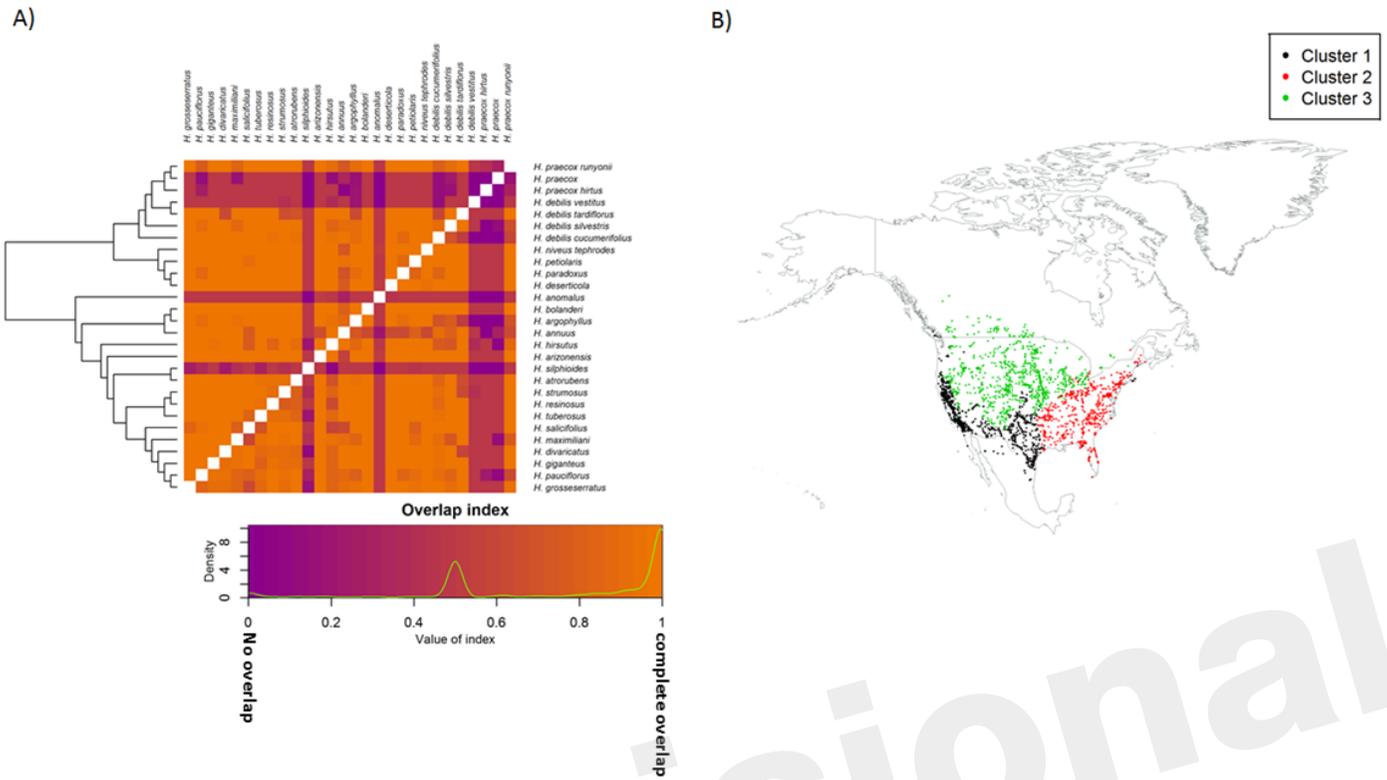
Figure 2. A) Map of North America showing the species richness of sunflower. B) Map of North America showing collection gaps for sunflower; in both maps lower numbers (bluish colors) indicates low species numbers and high numbers (reddish) indicate high species numbers in a given location, all areas colored require collection they differ only in the number of species that need to be collected within the geographic location.



Provisional

Figure 3.TIF

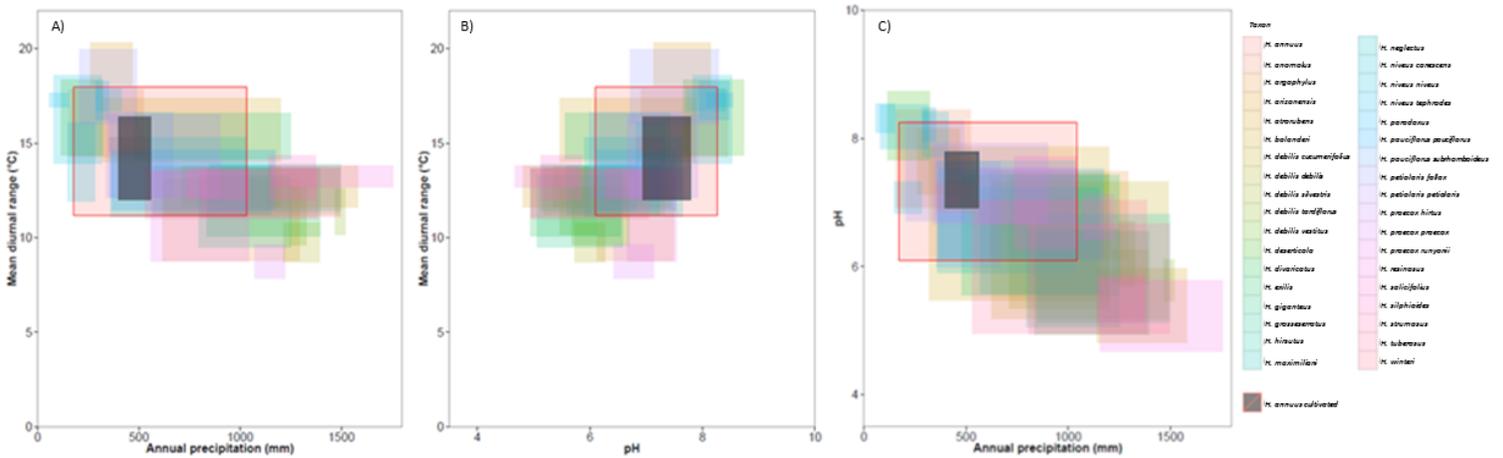
Figure 3. Geographic niche overlap based on bioclimatic and biophysical variables, both calculated by D (above diagonal) and I (below diagonal). Taxa are grouped by the phylogenetic relationship identified from the ETS sequences retrieved from NCBI. Values closer to 0 (no overlap = niche divergence) are purple while values closer to 1 (complete overlap = niche convergence) are orange; B) Occurrence points for each taxa grouped based on the first three principle components of biophysical and bioclimatic variables. Clusters share homogeneous bioclimatic and biophysical conditions.



Provisional

Figure 4.TIF

Figure 4. Climatic niches of *Helianthus* crop wild relatives for A) Mean diurnal range and annual precipitation, B) Soil pH and mean annual precipitation, C) Mean diurnal range and annual precipitation. Niches per taxa represent the middle 90% of occurrence points, i.e., 10% outliers are not included. Red boxes show the niche of wild *H. annuus* and black boxes show the niche of cultivated *H. annuus* in North America.



Provisional

Figure 5.TIF

Figure 5. Test of phylogenetic signal utilizing the K for 25 of 36 taxa analyzed with complete genetic and environmental information (Blomberg, et al, 2003). K measures phylogenetic signal in traits, where K values below 1 indicates low dependence of traits on evolutionary history (not conserved between taxa) and K values above 1 indicates trait conservation over evolutionary history (traits conserved over evolutionary time). *indicates K significantly greater than 1 ($p < 0.05$).

